

# NORTH PACIFIC SWORDFISH (Xiphiaus gladius) STOCK ASSESSMENT IN 2014 

## REPORT OF THE BILLFISH WORKING GROUP

International Scientific Committee for Tuna and Tuna-like Species in the North Pacific Ocean


16-22 July 2014
Taipei, Chinese-Taipei

## EXECUTIVE SUMMARY

Introduction: This Executive Summary updates assessment information for the two North Pacific swordfish stocks. The updated assessments were conducted by the Billfish Working Group of the International Scientific Committee on Tuna and Tuna-Like Species in the North Pacific. The Executive Summary summarizes assessment information on stock status relative to MSY-based reference points, stock projections, and conservation advice, as well as providing current information on stock identification and distribution, fishery catches, data and assessment, biological reference points, and special comments.

Stock Identification and Distribution: Swordfish (Xiphias gladius), also known as broadbill swordfish, inhabit a wide region of the Pacific between the latitudes of $50^{\circ} \mathrm{N}$ and $50^{\circ} \mathrm{S}$. Swordfish is a highly migratory species with high economic value in both commercial and recreational fisheries. In the North Pacific, the swordfish (Xiphias gladius) population is comprised of two stocks, separated by a diagonal boundary extending from Baja, California, to the Equator. These are the Western and Central North Pacific Ocean stock (WCNPO), distributed in the western and central Pacific, and the Eastern Pacific Ocean stock (EPO), distributed in the eastern Pacific (Figure 1).

Catches: For the WCNPO stock, time series of fishery catches by country show variability in swordfish yields over the past six decades (Figure 2.1). During the 1950s, Japanese distant-water and offshore longline fisheries accounted for more than $80 \%$ of the annual swordfish harvests. The total reported annual catch of WCNPO swordfish peaked at 22,000 metric tons in 1960 but rapidly decreased during the 1960s coincident with shifts in species targeting by longline fleets. During the 1970s, the average annual reported catch of swordfish in the WCNPO area was about 10,100 metric tons and the historically lowest catch of 6,800 metric tons occurred in 1972. Total annual swordfish catches increased slightly in the 1980s and reached a level of 15,800 metric tons in 1985 due to a few years of high catches by Japanese distant-water and offshore longline fleets and USA fisheries (Figure 2.1). Swordfish catch reached a high level of 19,000 mt in 1993 and declined to a low of $13,000 \mathrm{mt}$ in 1996. During the 2000s, the average annual reported catch of swordfish in the WCNPO was about 13,600 metric tons. After 2007, annual catches decreased substantially to average around 10,000 metric tons in 2011-2012.
In the EPO stock area, swordfish catches were low in the early years of the fishery and steadily increased until 1970, after which catch fluctuated between 2,000 and 7,500 mt through the 1990s (Figure 2.2). In 1998 and 2001-2002, annual catches were above $7,000 \mathrm{mt}$, and then declined to $3,235 \mathrm{mt}$ in 2006. Since then, catch has risen to an historic peak of $9,910 \mathrm{mt}$ in 2012 (Figure 2.2), averaging about $9,700 \mathrm{mt}$ in 2010-2012. In 2012, Japan, Spain, China, and Taiwan jointly caught $85 \%$ of the total swordfish harvest in the EPO.

Data and Assessment: For the WCNPO swordfish stock, catch data was updated for this assessment and this led to an increase of about $10 \%$ and $30 \%$ in reported catch biomass during 1960-1999 and 2000-2009, respectively. Fishery catch data were taken from all available fisherydependent data by Japan, Taiwan, Korea, USA, and other countries in the WCNPO stock area (Table 1 and Figure 2.1). Standardized fishery-dependent CPUE swordfish were estimated for Japanese distant water and offshore longline fisheries, Taiwanese distant water longline fisheries, and the shallow-set sector of the Hawaii-based pelagic longline.

Total catches of EPO swordfish from all countries and sources were updated during 1951-2012 (Table 2 and Figure 2.2) and recent catch data from 2007-2012 were recompiled using updated data provided by the IATTC, the WCPFC, and the individual countries of Japan, Taiwan, Korea, Mexico, and Chile (Figure 2.2). Estimates of standardized commercial fishery CPUE for EPO swordfish were provided by Japan and Taiwan through 2012.

Generalized surplus production models used for updating the WCNPO and EPO swordfish assessments had a very similar structure to the previous assessment and were formulated as Bayesian state space models with explicit observation and process error terms. Exploitable biomass time series were estimated from the observed relative CPUE abundance indices and from catches using observation error likelihood function and prior distributions for model parameters. Parameter estimation was based on Markov Chain Monte Carlo simulation using Gibbs sampling was applied to numerically sample the posterior distribution of quantities of interest, e.g. exploitable biomass.

Table 1. Reported annual values of catch (mt) and posterior mean values of exploitable biomass ( $\mathrm{B}, \mathrm{mt}$ ), relative biomass ( $\mathrm{B} / \mathrm{B}_{\mathrm{MSY}}$ ), harvest rate (percent of exploitable biomass), relative harvest rate $\left(\mathrm{H} / \mathrm{H}_{\text {MSY }}\right)$, and probability of annual harvest rate exceeding $\mathrm{H}_{\text {MSY }}$ for the Western and Central North Pacific swordfish stock.

| Year | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | Mean ${ }^{1}$ | Min ${ }^{1}$ | Max ${ }^{1}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Reported Catch | 15,051 | 15,799 | 13,631 | 12,375 | 10,670 | 9,456 | 9,863 | 12,962 | 6,753 | 21,972 |
| Exploitable Biomass | 76,320 | 72,290 | 68,620 | 68,770 | 68,970 | 68,560 | 72,500 | 81,860 | 60,200 | 121,300 |
| Relative Biomass | 1.26 | 1.19 | 1.13 | 1.13 | 1.14 | 1.13 | 1.20 | 1.35 | 0.99 | 2.00 |
| Harvest Rate | 21\% | 23\% | 21\% | 19\% | 16\% | 15\% | 14\% | 17\% | 10\% | 31\% |
| Relative Harvest Rate | 0.84 | 0.93 | 0.84 | 0.76 | 0.66 | 0.59 | 0.58 | 0.69 | 0.39 | 1.23 |
| $\operatorname{Pr}\left(\mathrm{H}>\mathrm{H}_{\text {MSY }}\right)$ | 0.18 | 0.34 | 0.19 | 0.09 | 0.02 | 0.00 | 0.00 | 0.12 | 0.00 | 0.80 |

${ }^{1}$ During 1951-2012
Table 2. Reported annual values of catch (mt) and posterior mean values of exploitable biomass ( $\mathrm{B}, \mathrm{mt}$ ), relative biomass ( $\mathrm{B} / \mathrm{B}_{\mathrm{MSY}}$ ), harvest rate (percent of exploitable biomass), relative harvest rate $\left(\mathrm{H} / \mathrm{H}_{\mathrm{MSY}}\right)$, and probability of annual harvest rate exceeding $\mathrm{H}_{\text {MSY }}$ for the Eastern Pacific swordfish stock.

| Year | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | $\operatorname{Mean}^{1}$ | $\operatorname{Min}^{1}$ | $\operatorname{Max}^{1}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Reported Catch | 3,235 | 3,701 | 4,262 | 7,473 | 9,631 | 9,586 | 9,910 | 3,561 | 1 |  |
| Exploitable Biomass | 43,100 | 47,980 | 53,840 | 60,570 | 62,120 | 60,810 | 58,590 | 48,875 | 31,510 | 6,910 |
| Relative Biomass | 1.38 | 1.54 | 1.73 | 1.95 | 2.00 | 1.95 | 1.87 | 1.58 | 1.02 |  |
| Harvest Rate | $8 \%$ | $9 \%$ | $9 \%$ | $14 \%$ | $17 \%$ | $18 \%$ | $19 \%$ | $8 \%$ | $<1 \%$ |  |
| Relative Harvest Rate | 0.49 | 0.50 | 0.51 | 0.80 | 1.00 | 1.03 | 1.11 | 0.49 | 0.00 |  |
| Pr $\left(H>H_{\text {MSY }}\right.$ ) | 0.01 | 0.02 | 0.02 | 0.20 | 0.44 | 0.47 | 0.55 | 0.11 | 0.00 | 0.71 |

${ }^{1}$ During 1951-2012

Status of Stock: Exploitable biomass of WCNPO swordfish fluctuated at or above B $\mathrm{BSY}_{\mathrm{MS}}$ throughout the assessment time horizon and has remained high in recent years (Table 1 and Figure 3.1). As expected, there was an inverse pattern between estimated biomass and harvest rate as harvest rate fluctuated at or below $\mathrm{H}_{\text {MSY }}$. Trends in exploitable biomass and harvest rate from the current assessment are very similar to those from the 2009 assessment. In recent years, catches and harvest rates of WCNPO swordfish have had a declining trend, with exploitable biomass fluctuating around $70,000 \mathrm{mt}$, since 2007 (Table 1 and Figure 3.1). The Kobe plot showed that the WCNPO swordfish stock does not appear to have been overfished or to have experienced overfishing throughout most of the assessment time horizon of 1951-2012 (Figure 4.1). For the current status, results indicated it was very unlikely that the WCNPO swordfish population biomass was below $\mathrm{B}_{\mathrm{MSY}}$ in $2012\left(\operatorname{Pr}\left(\mathrm{~B}_{2012}<\mathrm{B}_{\mathrm{MSY}}\right)=14 \%\right)$. Similarly, it was extremely unlikely that the swordfish population was being fished in excess of $\mathrm{H}_{\mathrm{MSY}}$ in 2012 $\left(\operatorname{Pr}\left(\mathrm{H}_{2012}>\mathrm{H}_{\mathrm{MSY}}\right)<1 \%\right)$. Retrospective analyses indicated that there was no retrospective pattern in the estimates of exploitable biomass and harvest rate (Figure 5.1).
For the EPO stock, time series of estimates of exploitable biomass and harvest rate over the assessment time horizon differed from the previous assessment in recent years but have remained high in recent years (Table 2 and Figure 3.2). Exploitable biomass had a declining trend during 1969-1995 and has increased from 31,000 mt in 1995 to over $60,000 \mathrm{mt}$ in 2010, generally remaining above $\mathrm{B}_{\mathrm{MSY}}$. Harvest rates were initially low, have had a long-term increasing trend, and likely exceeded $\mathrm{H}_{\text {MSY }}$ in 1998, 2002, 2003, and also the most recent year, 2012 (Figure 3.2). The Kobe plot showed that overfishing likely occurred in only a few years, but may be occurring in recent years (Figure 4.2). In 2012, there was a $55 \%$ probability that overfishing was occurring in 2012, but there was a less than $1 \%$ probability that the stock was overfished. Retrospective analyses indicated that there was a clear retrospective pattern of underestimating exploitable biomass and overestimating harvest rate (Figure 5.2).

Projections and Risk Analyses: For the WCNPO stock, stochastic projections for eight harvest scenarios were conducted through 2016 (Figure 6.1). Results relative to MSY-based reference points indicated that exploitable biomass would likely remain above $\mathrm{B}_{\text {MSY }}$ through 2016 under the status quo catch or status quo harvest rate scenarios (Figure 6.1). For the high harvest rate scenarios (i.e., Maximum observed harvest rate, $150 \%$ of $\mathrm{H}_{\text {MSY }}, 125 \%$ of $\mathrm{H}_{\text {MSY }}$ ), exploitable biomass was projected to decline below $\mathrm{B}_{\text {MSY }}$ by 2016 (Figure 6.1) with harvest rates exceeding $\mathrm{H}_{\text {MSY }}$. In comparison, the stock would not be expected to experience any overfishing during 2014-2016 under the status quo catch and status quo harvest rate scenarios. (Figure 6.1) The risk analyses of harvesting a constant annual catch of WCNPO swordfish during 2014-2016 showed that there would be virtually no chance of the stock being overfished or experiencing overfishing in 2016 (Figure 7) if current annual catches of about $10,000 \mathrm{mt}$ were maintained. Annual catches of WCNPO swordfish would need to increase to roughly $15,000 \mathrm{mt}$ to have a moderate ( $50 \%$ chance) risk of overfishing and would need to increase to over $25,000 \mathrm{mt}$ to exceed a moderate risk of the stock being overfished in 2016 (Figure 7).

For the EPO stock, stochastic projections showed that exploitable biomass will likely have a decreasing trajectory during 2014-2016 under all eight of the harvest scenarios examined (Figure 6.2). Under the high harvest rate scenarios (status quo catch, Maximum observed harvest rate, $150 \%$ of $\mathrm{H}_{\text {MSY }}$ ), exploitable biomass was projected to decline to be roughly equal to $\mathrm{B}_{\text {MSY }}$ in 2016 (Figure 6.2) and maintain harvest rates above $\mathrm{H}_{\text {MSY }}$. In comparison, under the status quo
harvest rate scenario, exploitable biomass was projected to decline to only $40,000 \mathrm{mt}$ by 2016 , well above the $\mathrm{B}_{\mathrm{MSY}}$ level. Overall, the projections showed that if recent high catch levels persist, exploitable biomass will very likely decrease and a moderate risk of overfishing will likely continue to occur. The risk analyses for harvesting a constant catch of EPO swordfish during 2014-2016 showed that the probabilities of overfishing and becoming overfished increased as projected catch increased in the future (Figure 7). Maintaining the current catch of EPO swordfish of approximately $9,700 \mathrm{mt}$ would lead to a moderate risk of overfishing in 2016 but would lead to less than $1 \%$ probability of the stock being overfished in 2016.

Biological Reference Points: Biological reference points based on maximum sustainable yield were calculated from the generalized surplus production model results for the WCNPO and EPO swordfish stocks (Table 3). For WCNPO swordfish (Table 3), the point estimate and coefficient of variation (CV) of maximum sustainable yield, exploitable biomass to produce MSY, and harvest rate to produce MSY were: MSY $=14.92$ thousand mt with $\mathrm{CV}=12 \%, \mathrm{~B}_{\text {MSY }}=60.72$ thousand mt with $\mathrm{CV}=19 \%$, and $\mathrm{H}_{\mathrm{MSY}}=0.25$ with $\mathrm{CV}=22 \%$.
For EPO swordfish (Table 3), the point estimate and CV of maximum sustainable yield, exploitable biomass to produce MSY, and harvest rate to produce MSY were: MSY = 5.49 thousand mt with $\mathrm{CV}=30 \%$, $\mathrm{B}_{\text {MSY }}=31.17$ thousand mt with $\mathrm{CV}=22 \%$, and $\mathrm{H}_{\text {MSY }}=0.18$ with $\mathrm{CV}=34 \%$. Overall, the biological reference points indicated that the WCNPO stock was larger and more productive than the EPO stock.
Table 3. Estimates of current levels of exploitable biomass ( $\mathrm{B}_{2012}$, thousand mt ), average harvest rate ( $\mathrm{H}_{2010-2012}$, percent of exploitable biomass), and recent average yield ( $\mathrm{C}_{2010-2012}$, thousand mt ) along with estimated MSY-based biological reference points for the WCNPO and EPO swordfish stocks.

| Reference Point | WCNPO Stock Estimate | EPO Stock Estimate |
| :---: | :---: | :---: |
| $\mathrm{B}_{2012}$ | $72,500 \mathrm{mt}$ | $58,590 \mathrm{mt}$ |
| $\mathrm{H}_{2010-2012}$ | $15 \%$ | $18 \%$ |
| $\mathrm{C}_{2010-2012}$ | $9,996 \mathrm{mt}$ | $9,709 \mathrm{mt}$ |
| $\mathrm{B}_{\mathrm{MSY}}$ | $60,720 \mathrm{mt}$ | $31,170 \mathrm{mt}$ |
| $\mathrm{H}_{\mathrm{MSY}}$ | $25 \%$ | $18 \%$ |
| MSY | $14,920 \mathrm{mt}$ | $5,490 \mathrm{mt}$ |

Conservation Advice: Based on the assessment update, the WCNPO swordfish stock is not currently overfished and is not experiencing overfishing. The WCNPO stock is not fully exploited.

For the EPO swordfish stock, results indicated that overfishing may be occurring in recent years, and the recent average yield of roughly $10,000 \mathrm{mt}$, or almost two times higher than the estimated

MSY, is not likely to be sustainable in the long term. While biomass of the EPO stock appears to be nearly twice BMSY, any increases in catch above recent ${ }^{1}$ levels should consider the uncertainty in stock structure and unreported catch.

Special Comments: The WG recognized unreported catch and stock structure as two potential sources of uncertainty that were not accounted for in the stock assessments and either source would increase the overall uncertainty in the assessment results.
${ }^{1}$ recent is 3-year average for 2010-2012.

Figure 1. Two-stock structure for swordfish (Xiphias gladius) in the North Pacific Ocean, indicating separate stocks in the Western and Central Pacific Ocean and in the Eastern Pacific Ocean.


Figure 2.1 Swordfish (Xiphias gladius) catch (metric tons) in the Western and Central North Pacific Ocean stock area from 1951-2012 by country. †Other: catch data from Belize, Cook Islands, China, Spain, Fiji, Federated States of Micronesia, Kiribati, Marshall Islands, Papua New Guinea, Senegal, Tuvalu, and Vanuatu.


Figure 2.2. Swordfish (Xiphias gladius) catch (metric tons) in the Eastern Pacific Ocean stock area from 1951-2012 by country. $\dagger$ Other: catch data from Belize, Cook Islands, China, Spain, Fiji, Federated States of Micronesia, Kiribati, Marshall Islands, Papua New Guinea, Senegal, Tuvalu, and Vanuatu.


Figure 3.1. Trends in exploitable biomass (top) and harvest rate (bottom) of swordfish (Xiphias gladius) in the Western and Central North Pacific Ocean stock area. Estimated mean values from the posterior distribution (black circles and solid line), $95 \%$ confidence interval bars (solid vertical lines), and estimated biological reference points ( $\mathrm{B}_{\mathrm{MSY}}$ and $\mathrm{H}_{\mathrm{MSY}}$, horizontal dashed lines) are presented.


Figure 3.2. Trends in exploitable biomass (top) and harvest rate (bottom) of swordfish (Xiphias gladius) in the Eastern Pacific Ocean stock area. Estimated mean values from the posterior distribution (black circles and solid line), $95 \%$ confidence interval bars (solid vertical lines), and estimated biological reference points ( $\mathrm{B}_{\mathrm{MSY}}$ and $\mathrm{H}_{\text {MSY }}$, horizontal dashed lines) are presented.


Figure 4.1. Kobe diagram showing the estimated trajectories of relative exploitable biomass ( $\mathrm{B} / \mathrm{B}_{\mathrm{MSY}}$ ) and relative harvest rate $\left(\mathrm{H} / \mathrm{H}_{\mathrm{MSY}}\right)$ for swordfish (Xiphias gladius) in the Western and Central North Pacific Ocean stock area during 1951-2012.


Figure 4.2. Kobe diagram showing the estimated trajectories of relative exploitable biomass $\left(\mathrm{B} / \mathrm{B}_{\mathrm{MSY}}\right)$ and relative harvest rate $\left(\mathrm{H} / \mathrm{H}_{\mathrm{MSY}}\right)$ for swordfish (Xiphias gladius) in the Eastern Pacific Ocean stock area during 1951-2012.


Figure 5.1. Retrospective analyses of the absolute changes in exploitable biomass (a) and harvest rate (b) for swordfish (Xiphias gladius) in the Western and Central North Pacific stock area based on successive removals of one-year of assessment data and refits of the baseline production model.


Figure 5.2. Retrospective analyses of the absolute changes in exploitable biomass (a) and harvest rate (b) for swordfish (Xiphias gladius) in the Eastern Pacific stock area based on successive removals of one-year of assessment data and refits of the baseline production model.


Figure 6.1. Stochastic projections of expected exploitable biomass ( 1000 metric tons) of swordfish (Xiphias gladius) in the Western and Central Pacific Ocean stock area during 20132016 under alternative harvest rates. Upper panel shows projection results of applying a harvest rate set to be $50 \%, 75 \%, 100 \%, 125 \%$, and $150 \%$ of the value of estimate of $\mathrm{H}_{\mathrm{MSY}}$ (denoted as $F_{\text {MSY }}$ in the Figure). Lower panel shows projection results of applying a status quo harvest rate based on the 2010-2012 average estimates, a status quo catch based on the 2010-2012 average catch, and the maximum observed harvest rate in the 1951-2012 time series.


Figure 6.2. Stochastic projections of expected exploitable biomass ( 1000 metric tons) of swordfish (Xiphias gladius) in the Eastern Pacific Ocean stock area during 2013-2016 under alternative harvest rates. Upper panel shows projection results of applying a harvest rate set to be $50 \%, 75 \%, 100 \%, 125 \%$, and $150 \%$ of the value of estimate of $\mathrm{H}_{\mathrm{MSY}}$ (denoted as $F_{\text {MSY }}$ in the Figure). Lower panel shows projection results of applying a status quo harvest rate based on the 2010-2012 average estimates, a status quo catch based on the 2010-2012 average catch, and the maximum observed harvest rate in the 1951-2012 time series.


Figure 7. Probabilities of experiencing overfishing ( $\mathrm{H}>\mathrm{HMSY}$, solid line), of exploitable biomass falling below BMSY ( $\mathrm{B}<0.5^{*}$ BMSY, open circles), and of being overfished relative to a reference level of $1 / 2 \mathrm{BMSY}$ ( $\mathrm{B}<0.5^{*}$ BMSY, solid squares) in 2016 for swordfish in the Western and Central Pacific Ocean stock area (a) and Eastern Pacific Ocean stock area (b) based on applying a constant catch biomass ( x -axis, thousand mt ) in the stock projections.



## INTRODUCTION

Swordfish (Xiphias gladius), also known as broadbill swordfish, inhabit a wide region of the Pacific between the latitudes of $50^{\circ} \mathrm{N}$ and $50^{\circ} \mathrm{S}$ (Ward and Elscot, 2000). Swordfish is a highly migratory species with high economic value in both commercial and recreational fisheries. In the North Pacific, the majority of catch has been taken by longline fishing vessels from Japan, Taiwan and the United States, which accounted for $95 \%$ of the total harvest in the North Pacific in 2010s, with the remaining catch taken by China, Korea, Mexico, and Spain.

Several stock structures have been proposed for Pacific swordfish (Alvarado Bremer et al., 2006; Ichinokawa and Brodziak, 2008). Stock assessments on swordfish in the North Pacific have been conducted primarily using catch, and abundance indices in the form of catch-per-unit effort, or CPUE. In 2004, Kleiber and Yokawa (2004) used MULTIFAN-CL to assess North Pacific swordfish in a four-region model. It has been suggested that these model fits and parameter estimates were sensitive to model structure. In two subsequent studies, a similar lengthstructured modeling approach was applied, which included some sex-specific data (Wang et al. 2005, 2007). These previous studies concluded that there was little contrast in the North Pacific swordfish fishery CPUE data to estimate stock status relative to biological reference points. Updated catch and effort data, however, were expected to improve model fits and to help estimate recent trends in swordfish abundance and harvest rates.

In 2009, all swordfish in the North Pacific were assessed as both a single stock north of the Equator and also under a two-stock scenario, with one stock in the Western and Central Pacific Ocean (WCNPO) and another in the Eastern Pacific Ocean (EPO) (ISC 2009), separated by a diagonal boundary extending from Baja, California, to the Equator (
Figure 2), based on the analysis by Ichinokawa and Brodziak (2008). The EPO swordfish stock assessment was revised in 2010 using updated catch data (Brodziak 2010). The previous assessment results indicated that for both swordfish stocks, current biomasses were above the biomass at which the maximum sustainable yield (MSY), or maximum surplus production would be obtained and harvest rates were below the harvest rate to produce MSY (ISC, 2009).

Based on the scientific consensus that a two-stock scenario is likely, we present here an updated assessment of the WCNPO and EPO swordfish stocks. The WCNPO stock is distributed in the North Pacific Ocean west of a diagonal boundary that extends from $170{ }^{\circ} \mathrm{W}$ towards Baja California (Figure 1) (Ichinokawa and Brodziak 2008). The EPO swordfish stock is centered on the Equator in the Eastern Pacific, bounded on the south by $20^{\circ} \mathrm{S}$ and extending northeast diagonally from $170^{\circ} \mathrm{W}$ towards Baja California (
Figure 2).
We applied a Bayesian statistical framework to estimate parameters of production models to assess the swordfish population in the WCNPO area using updated catch and effort through 2012. The Bayesian method provided direct estimates of parameter uncertainty that were straightforward to interpret and were appropriate for risk analyses. The production models include both process error for biomass production dynamics and observation errors for fitting the observed CPUE data from multiple fishing fleets. The assessment model estimated biological reference points, biomass, harvest rate, stock status, and associated uncertainties. The objectives
of this study are: (i) to update the ISC (2009) stock assessment for the WCNPO and EPO stocks, (ii) to develop Bayesian posterior distributions for quantities of management interest using Markov chain Monte Carlo (MCMC) simulation, (iii) to examine the sensitivity of the results of the assessment to changes to its prior assumptions, (iv) to conduct a retrospective analysis of stock assessment estimates, and (v) to conduct future stock projections accounting for uncertainty in stock size estimates and process error.

## MATERIALS AND METHODS

## Fishery Data

## Catch

For the WCNPO swordfish stock, fishery catch data by country from 1951-2012 for assessing WCNPO swordfish were taken from the most recent summary of available fishery-dependent data (Kimoto and Yokawa, 2014; Ito and Childers, 2014). Commercial catch of swordfish caught by Japan, Taiwan, Korea, USA, and other countries in the WCNPO stock area were updated from the 2009 assessment (Table 1.1, Figures 1 and 2.1). More specifically, Japan, Taiwan, Korea, and the USA directly provided updated catch data, and swordfish catches for all other fishing countries in the WCNPO area were collected from WCPFC 2005-2012 and IATTC 20072012 category II data (Tagami et al., 2014, Figure 2.1). Japanese swordfish fishery data included Japanese coastal, offshore, and distant-water longliners and other coastal gears. Taiwanese swordfish fishery data included the distant water longline, offshore longline and costal fisheries while Korean swordfish fishery data included distant water longline fishery. For the IATTC swordfish fishery data, the swordfish catch numbers in WCNPO area were converted to catch biomass by using the annual averaged weight that derived from the size-frequency data and the relationship between body biomass ( $\mathrm{W}, \mathrm{kg}$ ) and eye-fork-length (EFL, cm) (DeMartini et al. 2000, DeMartini et al. 2007, Uchiyama and Humphreys, 2007):

$$
\begin{equation*}
W=0.0000137 * L^{3.04} \tag{1}
\end{equation*}
$$

where $W$ is weight in kg and $L$ is eye-fork length in cm . For the WCPFC swordfish fishery data, swordfish catch biomasses were also separated by stock area, and the WCNPO stock included catch data from Belize, Cook Islands, China, Spain, Fiji, Federated States of Micronesia, Kiribati, Marshall Islands, Papua New Guinea, Senegal, Tuvalu, and Vanuatu (Table 1.1).

Swordfish are also mostly caught by longline fisheries in the EPO, some of which target other pelagic species such as tuna. The annual EPO swordfish catch has fluctuated between 3,000 to almost 10,000 metric tons ( mt ) since 2000 . The majority of catch has been taken by longline fishing vessels from Japan, Spain, China, Korea, and Taiwan (Table 1.2, Figures 1 and 2.2), which accounted for $91 \%$ of the total harvest in the Eastern Pacific in 2012. The remaining catch was taken by Belize, Mexico, Chile, French Polynesia, Peru, Vanuatu, and the United States.

Fishery catch data for swordfish in the Eastern Pacific Ocean from 1951-2012 were compiled from several sources. Catch data from 1951-2006 were taken from the most recent summary of available fishery-dependent data during the previous assessment (Brodziak 2010). More recent catch data from 2007-2012 were compiled using data provided by the Inter-American Tropical Tuna Commission (IATTC), Western and Central Pacific Fisheries Commission (WCPFC), and individual countries of Japan, Taiwan, Korea, Mexico, and Chile (Table 1.2 and Figure 2.2).

When a country provided catch data directly to the ISC Billfish Working Group, those data were considered more accurate and were used in lieu of data reported to the IATTC and WCPFC. Overall, the catch data were used to model the effects of fishery removals from the EPO swordfish stock during 1951-2012. A description of each dataset follows.

The IATTC provided a catch dataset for 2007-2012 describing total numbers of swordfish caught by longline by year, country, latitude, and longitude. The IATTC also provided a separate smaller dataset on lengths, indicating the total numbers of swordfish caught and their sizes in cm by year, country, latitude, and longitude. Each dataset was separated into data for the EPO stock and for the WCNPO stock. The lengths dataset was used to convert total numbers caught in the catch dataset into biomass. First, the lengths were converted into biomass using the lengthweight relationship for swordfish in Eqn (1). From these weights, the average weight of a swordfish caught in each year was calculated, and this average yearly weight was used to convert the numbers of swordfish caught in the catch dataset into swordfish catch biomass. The catch dataset was then aggregated by country and year to calculate the annual swordfish catch biomass.

Based on fishery information from the IATTC, the entire longline swordfish catch of Peru was expected to have been harvested in the EPO. As a result, the entire Peruvian catch time series from 1954-2010 was added to the catch data for swordfish in the EPO. Catch data from Peru came from the most recent assessment of swordfish conducted by the IATTC (Hinton and Maunder 2011). This is the first time that swordfish data from Peru were included in the EPO assessment. The annual EPO swordfish catch from Peru during 2011-2012 was estimated as the average catch from 2007-2010.

Similarly, the WCPFC provided data for 2007-2012 north of the Equator on swordfish numbers and tons caught by year, country, latitude, and longitude. These data were separated by stock area (EPO versus WCNPO) and were aggregated by country and year to calculate the total tons of swordfish caught by each country in each year (Tables 1.1 and 1.2).

Swordfish catches for Japan, Taiwan, Korea, Mexico, and Chile were collected from the individual countries. Japan provided total swordfish catch in mt from their offshore and distantwater longline fleet for 1951-2012, with data from 2011 and 2012 still preliminary (Kimoto and Yokawa 2014). These data were used for 2007-2012, since it was considered more accurate than the reported IATTC and WCPFC data. The updated Japanese catch data for 1951-2006 were considered the best available data to date, and as a result, Japanese catch time series used in the previous assessment was replaced with the updated data and the total catch biomass time series was updated. Taiwan provided total swordfish catch biomass from their offshore and distant water longline fleet from 1964-2012. The updated Taiwanese catch data for 1964-2006 were used in lieu of the Taiwanese time series of catch used in the previous assessment. Taiwan also provided a brand new time series of swordfish catch for their offshore longline and other fisheries. The total catch of swordfish in the EPO was updated using these two catch time series. In particular, the Taiwanese catch data for 2007-2012 were used in place of Taiwanese data reported to IATTC and WCPFC, which included some minor differences. Korea provided total swordfish catch biomass for 2007-2012 from their tuna longline fisheries, by year, latitude, and longitude. These data was separated by stock area (EPO versus WCNPO), and then aggregated by year to calculate the Korean swordfish catch biomass by year during 2007-2012. Again, these
country-specific data were used in lieu of catch data for Korean data reported to the IATTC and WCPFC, which included some minor differences.

Swordfish catch biomass data for the Mexican longline fishery during 2007-2010 were taken from the most recent ISC country report submitted by Mexico (Dreyfus et al. 2013). The annual EPO swordfish catch for Mexico during 2011-2012 was estimated as the average annual catch during 2007-2010. Data were not available by latitude and longitude, but catch distribution maps indicated that the vast majority of swordfish were caught in the EPO rather than in the WCNPO. As a result, all swordfish caught by Mexico were assumed to be from the EPO stock. Swordfish catch data from Chile was updated for 2007-2012 using Annual Statistics of Fisheries and Aquaculture reports from the Chilean fisheries agency, Servicio Nacional de Pesca y Acuicultura (SERNAPESCA 2007-2012). At the guidance of the IATCC, it was assumed that swordfish landed in Chile's two northernmost fishery regions (Regions XV and I), which lie north of the southern boundary of the EPO, were likely harvested in the EPO. The total landings of swordfish from these two regions were added to the total EPO catch by year.

## Catch-Per-Unit Effort

For WCNPO swordfish, standardized fishery-dependent CPUE swordfish were estimated for Japanese distant water and offshore longline fisheries, Taiwanese distant water longline fisheries, and the shallow-set sector of the Hawaii-based pelagic longline fishery (Table 2.1, Figures 1 and 3.1). In particular, monthly aggregated dataset by $5 \times 5$ degree grids from 1952-1974 and those gear configurations from 1975 to 2012 were used in the CPUE standardization for Japanese distant water and offshore longline fisheries (Kimoto et al., 2014). The two standardized CPUE series were combined into a single period from 1952-2012 $(\mathrm{n}=61)$ using the average ratio of the standardized CPUEs for the time period of series overlap between 1975 and 1979 (Table 2.1 and Figure 3.1). Alternative CPUE series without Japanese designated areas 8 and 9 were also provided by Kimoto et al. (2014).

For Taiwanese distant water longline fisheries, aggregated data by $5 \times 5$ degree grids, month, and gear configurations were used for CPUE standardization (Sun et al., 2014). Information about gear configuration was only available since 1995. It was noted that there has been a change in target species and fishing grounds in this fishery around 2000. To account for this change, two standardized CPUE series for the separate time periods of 1969-1999 $(\mathrm{n}=25)$ with several missing values and 2000-2012 $(\mathrm{n}=13)$ were developed (Table 2.1 and Figure 3.1).

Operational data in the shallow-set sector of the Hawaii-based pelagic longline fishery in 19952012 collected by fishery observers were used for CPUE standardization (Walsh and Brodziak, 2014). Swordfish is the target species in the shallow-set fishery sector, which was closed between 2001 and 2004 due to fishery interactions with protected sea turtles. Because of this temporal gap, the CPUE standardization analyses used data from 1995-2000 $(\mathrm{n}=6)$ and 2005-2012 $(\mathrm{n}=8)$ to estimate standardized shallow-set CPUE (Table 2.1 and Figure 3.1).

For EPO swordfish, estimates of standardized commercial fishery CPUE were provided by Japan (Kimoto et al. 2014) and Taiwan (Sun et al. 2014) through 2012 (Table 2.2, Figures 1 and 3.2). The Japanese longline CPUE time series spanned 58 years (1955-2012), but was divided into three separate series: 1955-1974, 1975-1993, and 1994-2012 (Table 2.2 and Figure 3.2). The

Taiwanese distant water longline CPUE time series spanned 13 years (2000-2012) and included information on hook per float in the CPUE standardization (Table 2.2 and Figure 3.2). A second Taiwanese distant water longline CPUE time series exists for 1968-1999, but ultimately was not used because the inclusion of this CPUE series resulted in a lack of MCMC convergence and a very poor fit to CPUE data. The standardized CPUE series from Japan and Taiwan served as relative abundance indices for swordfish in the EPO, and were used to model changes in the relative abundance of swordfish through time. We calculated the Pearson correlation coefficient for the two CPUE series that overlapped in time: Japanese CPUE from 1994-2012, and Taiwanese CPUE from 2000-2012. The relative CVs of CPUE were all assumed to be a value of 1, that is, annual observation error variances were set to be equal for each CPUE value in a time series (Brodziak 2010).

## Bayesian Production Model

## Biomass Dynamics

Swordfish production models followed a similar structure to the previous production model used for Pacific swordfish ( Meyer and Millar, 1999; Brodziak and Ishimura 2009; Brodziak 2010). Production models were formulated as Bayesian-state space models with explicit observation and process error terms (e.g., Meyer and Millar 1999, Brodziak 2007). We implemented the state-space models in WinBUGS (version 1.4.3, Lunn et al. 2000) via the R2WinBUGS package (Sturtz et al., 2005) in the statistical programming environment R (R Development Core Team 2008). The biomass time series comprised the unobserved state variables which were estimated from the observed relative abundance indices (i.e., CPUE) and from catches using observation error likelihood function and prior distributions for model parameters $(\theta)$. In this case, the observation error likelihood measured the discrepancy between observed and predicted CPUE, and the prior distributions represented the relative degree of belief about the possible values of model parameters.

The process dynamics represented the fluctuations in exploitable swordfish biomass due to density-dependent processes and fishery harvests. The biomass dynamics were based on a generalized production model with an annual time step. Under this three-parameter model, biomass in year $T\left(B_{T}\right)$ depends on the previous biomass ( $B_{T-1}$ ), catch ( $C_{T-1}$ ), intrinsic growth rate $(R)$, carrying capacity $(K)$, and a production shape parameter $(S)$ for $T=2, \ldots, N$ :

$$
\begin{equation*}
B_{T}=B_{T-1}+R \cdot B_{T-1}\left(1-\left(\frac{B_{T-1}}{K}\right)^{S}\right)-C_{T-1} \tag{2}
\end{equation*}
$$

The production model shape parameter, $S$, determines where surplus production peaks as biomass varies as a fraction of carrying capacity. If the shape parameter is less than unity ( $0<S<1$ ), then surplus production peaks when biomass is below $1 / 2$ of $K$ (i.e., a left-skewed production curve) and the stock has relatively high productivity. If the shape parameter is greater than unity $(S>1$ ), biomass production is highest when biomass is above $1 / 2$ of $K$ (i.e., a rightskewed production curve), and the stock has relatively low productivity. If the shape parameter is identically unity $(S=1)$, the production model is identical to a discrete-time Schaefer production model where maximum surplus production occurs when biomass is equal to $1 / 2$ of $K$. Thus, the shape of the biomass production curve can be symmetric, right-, or left-skewed depending on the estimated value of S .

The generalized production model was re-parameterized using the proportion of carrying capacity $(P=B / K)$ to improve the efficiency of the Markov Chain Monte Carlo algorithm used to estimate parameters (i.e., Meyer and Millar 1999). Given this parameterization, the process dynamics are:

$$
\begin{equation*}
P_{T}=P_{T-1}+R \cdot P_{T-1}\left(1-P_{T-1}^{S}\right)-\frac{C_{T-1}}{K} \tag{3}
\end{equation*}
$$

## Biological Reference Points

The values of biomass and annual harvest rate that maximize biomass production are relevant as biological reference points for maximum sustainable yield (MSY). For the generalized production model, the biomass that produced $M S Y\left(B_{\mathrm{MSY}}\right)$ is:

$$
\begin{equation*}
B_{M S Y}=K \cdot(S+1)^{\frac{-1}{S}} \tag{4}
\end{equation*}
$$

The corresponding annual harvest rate that produced $M S Y\left(H_{M S Y}\right)$ was:

$$
\begin{equation*}
H_{M S Y}=R\left(1-\frac{1}{S+1}\right), \tag{5}
\end{equation*}
$$

and the associated value of maximum sustainable yield (MSY) was:

$$
\begin{equation*}
M S Y=R\left(1-\frac{1}{S+1}\right) \cdot K(S+1)^{\frac{-1}{S}} \tag{6}
\end{equation*}
$$

Note that $H_{M S Y}$ can be converted to its instantaneous equivalent, $F_{M S Y}$ by the following equation:

$$
\begin{equation*}
F_{M S Y}=-\log \left(1-H_{M S Y}\right) . \tag{7}
\end{equation*}
$$

As a result, the generalized production model produced estimates of biological reference points for swordfish that can be directly used for determining stock status with respect to MSY-based reference points and this conservation information is provided in the current assessment.

## Observation Error Model

The observation error model relates the observed fishery CPUE to the exploitable biomass of the swordfish stock under each scenario. It is assumed that each CPUE index ( $I$ ) is proportional to biomass with catchability coefficient $Q_{I}$ :

$$
\begin{equation*}
I_{T}=Q_{I} B_{T}=Q_{I} K P_{T} \tag{8}
\end{equation*}
$$

The observed CPUE values are subject to natural sampling variation which is assumed to be lognormally distributed. The observation errors are distributed as $v_{T}=e^{V_{T}}$, where the $V_{T}$ are independent and identically distributed normal random variables with a mean of 0 and variance $\tau_{I}{ }_{I}$ for CPUE series $I$.

Given the lognormal observation errors, the observation equations for each CPUE series $I$ for each year indexed by $T=1, \ldots, N$ are:

$$
\begin{equation*}
I_{T}=Q_{I} K P_{T} \cdot v_{T} \tag{9}
\end{equation*}
$$

This specifies the general form of the observation error likelihood function $p\left(I_{T} \mid \theta\right)$ for each fishing fleet through time.

## Process Error Model

The process error model compares the dynamics of exploitable biomass to natural variability in demographic and environmental processes affecting the swordfish stock. The deterministic process dynamics (Eqn. 3) are subject to natural variation as a result of fluctuations in life history parameters, trophic interactions, environmental conditions and other factors. In this case, the process error represents the joint effects of a large number of random multiplicative events which combine to form a multiplicative lognormal process under the Central Limit Theorem. As a result, the process error terms are assumed to be independent and lognormally distributed random variables $\eta_{T}=e^{U_{T}}$ where the $U_{T}$ are normal random variables with mean 0 and variance $\sigma^{2}$.

Given the process errors, the state equations define the stochastic process dynamics by relating the unobserved biomass states to the observed catches and the estimated population dynamics parameters. Assuming multiplicative lognormal process errors, the state equations for the initial time period $(T=1)$ and subsequent periods $(T>1)$ are:

$$
\begin{align*}
& P_{1}=\eta_{1} \\
& P_{T}=\left(P_{T-1}+R \cdot P_{T-1}\left(1-P_{T-1}^{S}\right)-\frac{C_{T-1}}{K}\right) \cdot \eta_{T} \text { for } T>1 \tag{10}
\end{align*}
$$

These coupled state equations set the conditional prior distribution for the proportion of carrying capacity, $p\left(P_{T}\right)$, in each time period $T$, conditioned on the proportion in the previous period.

## Prior Distributions

Under the Bayesian estimation framework, prior distributions were employed to quantify existing knowledge, or the lack thereof, of the likely value of each model parameter. For the production model, the model parameters consisted of the carrying capacity ( $K$ ), the intrinsic growth rate $(R)$, the shape parameter $(S)$, the catchability coefficients $\left(Q_{I}\right)$, the process and observation error variances ( $\sigma^{2}$ and $\tau_{I}^{2}$ ), and the annual biomasses as a proportion of carrying capacity $(P)$. Auxiliary information was incorporated into the formulation of the prior distributions when it was available. Information about the prior distributions used in the production model analyses were summarized for WCNPO (Table 3.1) and EPO swordfish (Table 3.2) stocks and details of the prior distributions are described below.

## Prior for Carrying Capacity

The prior distribution for the carrying capacity $p(K)$ is a lognormal distribution with mean $\left(\mu_{K}\right)$ and variance $\left(\sigma_{K}^{2}\right)$ parameters:

$$
\begin{equation*}
p(K)=\frac{1}{\sqrt{2 \pi} K \sigma_{K}} \exp \left(-\frac{\left(\log K-\mu_{K}\right)^{2}}{2 \sigma_{K}^{2}}\right) . \tag{11}
\end{equation*}
$$

The variance parameter is set to achieve a coefficient of variation (CV) for $K$ of $50 \%$, e.g., $C V[K]=\left(\exp \left(\sigma_{K}^{2}\right)-1\right)^{\frac{1}{2}}=0.5$. The mean $K$ values for WCNPO and EPO swordfish were set at $150,000 \mathrm{mt}$ and $75,000 \mathrm{mt}$, respectively. These mean values were taken from the previous assessments and reflect the order of magnitude of exploitable biomass likely needed to support the observed fishery catches.

## Prior for Intrinsic Growth Rate

The prior distribution for intrinsic growth rate $p(R)$ is a lognormal distribution with mean $\left(\mu_{R}\right)$ and variance $\left(\sigma_{R}^{2}\right)$ parameters set to achieve a CV for $R$ of $50 \%$ :

$$
\begin{equation*}
p(R)=\frac{1}{\sqrt{2 \pi} R \sigma_{R}} \exp \left(-\frac{\left(\log R-\mu_{R}\right)^{2}}{2 \sigma_{R}^{2}}\right) . \tag{12}
\end{equation*}
$$

The mean $R$ parameter for both WCNPO and EPO stocks was set to be $\mu_{R}=0.5$. This mean value is slightly higher than the range of prior means of $(0.40,0.43)$ estimated for North and South Atlantic swordfish, respectively, based on an analysis of life history parameters (McAllister et al. 2000). A similar analysis using life history parameters for North Pacific swordfish and the mean generation time approach (see McAllister et al. 2001) suggested higher mean values of $R$ of approximately 0.9 to 1.0 were appropriate. This analysis assumed female growth and maturation from DeMartini et al. (2000) and DeMartini et al. (2007) and used five alternative natural mortality rate estimators (Hoenig, Alverson and Carney, Pauly, Beverton-Holt $2^{\text {nd }}$ invariant, and Lorenzen Tropical) from Brodziak (2009) to calculate five alternative estimates of $R$. The primary difference between the Atlantic and Pacific swordfish life history parameters was the value of natural mortality. McAllister et al. (2000) assumed a constant natural mortality rate of $M$ $=0.2$ for Atlantic swordfish, while the Pacific swordfish natural mortality rate was estimated to be $M \approx 0.35$, roughly $75 \%$ higher than the Atlantic swordfish value. While there is uncertainty about an appropriate prior mean for $R$, setting the prior mean to be $\mu_{R}=0.5$ with a CV of $50 \%$ allows sufficient flexibility to estimate the probable value of $R$ given the observed catch and CPUE data.

## Prior for Production Shape Parameter

The prior distribution for the production function shape parameter $p(S)$ is a gamma distribution with rate parameter $\lambda$ and shape parameter $k$ :

$$
\begin{equation*}
p(S)=\frac{\lambda^{k} S^{k-1} \exp (-\lambda S)}{\Gamma(k)} \tag{13}
\end{equation*}
$$

For both WCNPO and EPO stocks, the values of the rate and shape parameters are set to $\lambda=k=$ 2. This choice of parameters sets the mean of $p(S)$ to be $\mu_{S}=1$, which corresponds to the value of $S$ for the Schaefer production model. This choice also implies that the CV of the shape parameter prior is $71 \%$. In effect, the shape parameter prior is centered on the symmetric Schaefer model as the default with sufficient flexibility to estimate a nonsymmetrical production function if needed. Prior for Catchabilities
The prior for the catchability coefficients $p\left(Q_{I}\right)$ for a given fleet $I$ is chosen to be a diffuse inverse-gamma distribution with scale parameter $\lambda$ and shape parameter $k$ :

$$
\begin{equation*}
p\left(Q_{I}\right)=\frac{\lambda^{k} Q_{I}^{-(k+1)}}{\Gamma(k)} \exp \left(\frac{-\lambda}{Q_{I}}\right) \tag{14}
\end{equation*}
$$

For both the WCNPO and EPO stocks, the scale and shape parameters are set to be $\lambda=k=0.01$. This choice of parameters implies that $1 / Q_{I}$ has a mean of 1 and a variance of 100 and produces a relatively uninformative prior. Since $1 / Q_{I}$ is unbounded at $Q_{I}=0$, an additional numerical constraint that $Q_{I}$ be no smaller than 0.0001 is imposed for the Markov Chain Monte Carlo sampling.

## Priors for Process and Observation Error Variances

For both swordfish stocks, the priors for the process error variance $p\left(\sigma^{2}\right)$ and observation error variance $p\left(\tau^{2}{ }_{I}\right)$ for each fleet $I$ are chosen to be inverse-gamma distributions. The choice of an inverse gamma distribution implies that the associated prior for error precision $\left(\pi=1 / \sigma^{2}\right)$ was effectively $p(\pi) \propto \pi^{-1}$ which is the Jeffrey's prior for the precision parameter (Congdon 2001).
As a result, inferences based on the gamma assumption are scale invariant and are not affected by changing the scale of the variance parameter. For the process error variance prior, the scale parameter is set to $\lambda=4$ and the shape parameter is $k=0.1$. This choice of parameters produces an expected value of approximately $E\left[\sigma^{2}\right]=0.025$ with a CV of $16 \%$. Similarly, for the observation error variance prior, the scale parameter is set to $\lambda=2$ and the shape parameter is $k=$ 0.45 . This choice of parameters produced an expected value of approximately $E\left[\tau_{I}^{2}\right]=0.223$
with a CV of $50 \%$. Given these prior assumptions, the initial observation error variance is roughly threefold greater than the process error variance. Of course, the posterior means of the process and observation errors estimated from the MCMC sampling also depend on the model fits to the observed data.

## Priors for Proportions of Carrying Capacity

Prior distributions for the time series of the proportion of biomass to carrying capacity, $p\left(P_{T}\right)$, are lognormal distributions as specified in the process dynamics. For both stocks, the mean proportion of carrying capacity for the initial year of $1951\left(P_{1}\right)$ was set to be 0.9 . This corresponded to an assumption that the North Pacific swordfish population was lightly exploited and had biomass near its carrying capacity following a period of limited directed fishing during World War II. To be consistent with the previous stock assessment, a CV of $10 \%$ was used. However, an alternative model configuration that used a CV of $50 \%$ was tested for the WCNPO stock to understand the effect of higher variation.

## Posterior Distribution

The joint posterior distribution of the swordfish production model needs to be sampled to make inferences about estimates of the model parameters. Given the catch data and $J$ series of standardized CPUE data to comprise the model data $D$, the posterior distribution $p(\theta \mid D)$ is proportional to the product of the prior distributions and the CPUE likelihood via Bayes theorem:

$$
\begin{equation*}
p(\theta \mid D) \propto p(K) p(R) p(S) p(Q) p\left(\sigma^{2}\right) \prod_{t=1}^{N} p\left(P_{t}\right) \prod_{j \in J} p\left(\tau_{j}^{2}\right) \prod_{T=1}^{N} p\left(I_{j, T} \mid \theta\right) \tag{15}
\end{equation*}
$$

Parameter estimation for this nonlinear multi-parameter model is based on generating a large number of independent samples from the posterior distribution. In this case, the Markov Chain Monte Carlo (MCMC) simulation using Gibbs sampling is applied to numerically generate a sequence of samples from the posterior distribution (Gilks et al. 1996). The WINBUGS software (Spiegelhalter et al. 2003) is used to set the initial conditions, perform the MCMC calculations, and summarize the results.

Markov Chain Monte Carlo simulations are conducted by simulating three chains of samples for each model. Each model was run for 800,000 iterations, sampled with a thinning rate of 25 with a burn-in period of 200,000 for three chains for a total of 72,000 samples to generate the posterior distributions.

A key issue in applying MCMC methods is how to determine when random draws have converged to the posterior distribution. Convergence of the MCMC samples to the posterior distribution was checked by monitoring the trace and diagnosing the autocorrelation plot. Convergence of the MCMC simulations to the posterior distribution was also checked using the Geweke diagnostic (Geweke 1992), Gelman and Rubin diagnostic (Gelman and Rubin 1992), and the Heidelberger and Welch stationarity and half-interval test (Heidelberger and Welch 1983), as implemented in the R Language ( R Development Core Team 2013) using the CODA software package (Plummer et al. 2006). These convergence diagnostics were monitored for several key model parameters (intrinsic growth rate, carrying capacity, production function shape parameter, and catchability coefficients) to verify convergence of the MCMC chains to the posterior distribution.

## Model Diagnostics

Goodness-of-fit to CPUE was measured to compare alternative production models using model residuals, root mean-squared error (RMSE), and the correlation between observed and predicted CPUE. Model residuals for the CPUE series are the $\log$-scale observation errors $\varepsilon_{T}$ :

$$
\begin{equation*}
\varepsilon_{T}=\ln \left(I_{T}\right)-\ln \left(Q K P_{T}\right) . \tag{16}
\end{equation*}
$$

A nonrandom pattern in the residuals indicates that the observed CPUE did not conform to one or more model assumptions. The RMSE of the CPUE fit provides another goodness-of-fit diagnostic with lower RMSE indicating a better fit when comparing models with the same number of parameters. Similarly, a higher correlation between observed and predicted CPUE indicates a better model match to observed CPUE trend.

The goodness of fit among different models with same data structure was evaluated by Deviance information criterion (DIC) (Spiegelhalter et al., 2002). The standardized log-residuals from the CPUE fit were visually examined for time trends. The Shapiro-Wilk test (Shapiro and Wilk, 1965) was used to test the normality of the standardized log-residuals. The estimates of production model can be problematic when the data are not informative about whether the population has a high K and a low r or vice versa (Hilborn and Walters, 1992). The posterior correlation between model parameters was examined for the base-case model.

## Sensitivity Analyses

The sensitivity of model outputs to priors was tested by varying the initial prior means of four key parameters: $R$, intrinsic growth rate; $K$, carrying capacity; $S$, shape parameter; and $P_{1}$, initial proportion of biomass to carrying capacity. For each of these prior means, we varied the prior mean by $25 \%$ higher and $25 \%$ lower, and compared resulting model outputs. These were considered to be useful high and low bounds for understanding which parameter was most important for estimating outputs, and more importantly, whether assessment results were robust to a $25 \%$ change in an input prior.

## Retrospective Analyses

We tested for any possible retrospective pattern (systematic inconsistencies among our model estimates of biomass and harvest rate based on increasing periods of data) by sequentially removing the most recent year of data going back 7 years, re-analyzing the model, and comparing estimated biomass and harvest rates. The within-model retrospective analyses were
used to examine changes in the estimates of exploitable biomass. Mohn's (1999) rho statistic (rho) was calculated as:

$$
\begin{equation*}
r h o=\sum_{y=1}^{\text {npeels }} \frac{B_{Y-y, t i p}-B_{Y-y, \text { ref }}}{B_{Y-y, \text { ref }}} \tag{3}
\end{equation*}
$$

where $B$ denotes exploitable biomass, $y$ denotes year, npeels denotes the number of years that are dropped in successive fashion and the assessment rerun, $Y$ is the last year in the full time series, tip denotes the terminal estimate from an assessment with a reduced time series, and ref denotes the assessment using the full time series.

## Projections

Stochastic projections were conducted to show the probable changes in exploitable biomass and catch under various harvest scenarios, including scenarios requested by the Western and Central Pacific Fisheries Commission's $9^{\text {th }}$ session of the Northern Committee. The following harvest scenarios were projected 4 years forward from 2012, the most recent year included in the assessment, to 2016:
a) Status quo harvest rate from the most recent 3 years
b) Status quo catch from the most recent 3 years
c) The maximum observed annual harvest rate
d) Harvest rate set at multiples of $0.5,0.75,1.0,1.25$, and 1.5 of $F_{M S Y}$

Projected harvest rates were sampled from a normal distribution with a mean corresponding to each scenario harvest value, and the standard deviation of harvest or catch values for the most recent 3 years (scenarios a and b) or a standard deviation assumed to be $5 \%$ of the mean value (scenarios c and d). Projections included process error and uncertainty in parameter estimation. The initial conditions for the projections are based on the MCMC samples from the estimated posterior distribution of exploitable swordfish biomass in the most recent year.

## Risk Analyses

Risk analyses to show the odds of depletion and overfishing were conducted. In these analyses, we calculated the probability of becoming overfished ( $B<0.5^{*} B_{M S Y}$ ) and overfishing ( $H>H_{M S Y}$ ) given a range of different projected future total catch levels for each stock. We projected 5 years forward from 2012, the most recent year of full data available to be included in the assessment, using catch levels at fixed intervals from 0 mt to a maximum of $40,000 \mathrm{mt}$, which is approximately four times the most recent average catch. Projected catch was sampled from a normal distribution centered at the projected catch level with a standard deviation of the most recent 3 years of catch. The initial conditions for the projections were based on the MCMC samples from the estimated joint posterior distribution of exploitable swordfish biomass and all other parameters in the most recent year. As a result, each projection harvest scenario included parameter estimation uncertainty, which, in turn, was incorporated into the estimated probabilities of overfishing or becoming overfished.

## RESULTS

## Fishery Data

## Catch

For the WCNPO swordfish stock, the updated catch led to an increase of about $10 \%$ and $30 \%$ in the 1960-2000 and 2000-2009 reported swordfish catch biomass, respectively, compared to the 2009 assessment. Time-series of fishery catches by country showed variability in swordfish yields over the past six decades (Table 1.1 and Figure 2.1). During the 1950s, Japanese distantwater and offshore longline fisheries accounted for more than $80 \%$ of the annual swordfish harvests. The total reported annual catch of WCNPO swordfish peaked at 22,000 metric tons in 1960. In the following decade, however, these fleets rapidly expanded for targeting tunas, and swordfish catches rapidly decreased during the 1960s. During the 1970s, the average annual reported catch of swordfish in the WCNPO area was about 10,100 metric tons and the historical lowest catch of 6,800 metric tons occurred in 1972. The total swordfish catch slightly increased in the 1980s and reached a level of 15,800 metric tons in 1985 resulting from a few years of higher catch of Japanese distant-water and offshore longline fleets and other USA fisheries (Figure 2.1). The swordfish catches by Japanese distant-water and offshore longline fleets showed a declining trend since 1990. However, there was a steep increase in Hawaii-based longline catches during the early 1990s and total swordfish catch reached a high level of 19,200 metric tons, then declined to a level of 13,700 metric tons in 1996-1999 (Figure 2.1). During the 2000s, the average annual reported catch of swordfish in the WCNPO was about 13,600 metric tons. After 2007, the total catches decreased significantly to around 10,000 metric tons and maintained at that level in 2011-2012. It should be noted a large fraction ( $25 \%$ ) of the swordfish catch has been taken by the Taiwanese offshore longline and other fisheries during this period.

Total catches of EPO swordfish were tabulated from all countries and sources (Table 1.2 and Figure 2.2) from 1951-2012. Swordfish catches were low in the early years of the fishery and steadily increased until 1970, after which catch fluctuated between 2,000 and 7,500 mt through the 1990s (Figure 2.2). In 1998 and 2001-2002, annual catches were above 7,000 mt, and then declined to $3,235 \mathrm{mt}$ in 2006. Since then, catch has risen to an historic peak of 9,910 mt in 2012 (Figure 2.2).

For the EPO swordfish stock, Japan and Spain had the highest swordfish catch in recent years (2007-2012), each catching over 2,000 mt in 2012 (Table 2.1 and Figure 2.2). China and Taiwan also caught large amounts of swordfish, over $1,500 \mathrm{mt}$ in 2012. These four countries (Japan, Spain, China, and Taiwan) jointly caught $85 \%$ of the total swordfish harvest in the EPO in 2012. Korea, Belize, Mexico, and Chile caught moderate amounts of swordfish. French Polynesia, Peru, the United States, and Vanuatu caught nominal amounts of swordfish.

## Catch-Per-Unit Effort

For the WCNPO stock, time-series of abundance indices available for this assessment showed some similarities in trends (Figure 3.1). Visual examination of the four CPUE indices suggested a similar trend of low CPUE in the 1970s, high CPUE in the early 1990s, and declining CPUE in the recent years among the indices used. Outliers in 1976, 1990, and 1995 were found in the Taiwanese distant water longline CPUE. The relative CV for Japanese distant water and offshore
longline CPUE during 1952-1974 was larger than the CPUE values during 1975-2012. Higher relative CPUE was also observed in the earlier period of Taiwanese distant water longline (19691999) and the Hawaii longline during 1995-1999 (Table 2.1 and Figure 3.1). There were no strong correlations $(|\rho| \geq 0.5)$ between CPUE time series. All pairs of CPUE indices were weakly or moderately positively correlated and had Pearson correlations ranging from 0.17 to 0.3 , with the exception of the correlation between the Japanese distant water and offshore longline CPUE during 1952-2012 and the Taiwanese distant water longline CPUE during 2000-2012 ( $\rho=-0.06$ ) and the correlation between the Hawaii longline CPUE during 1995-2012 and the Taiwan DW longline CPUE during 1969-1999 ( $\rho=-0.22$ ).

For the EPO stock, the two early standardized CPUE time series for Japan are each relatively stable, fluctuating around an average value (Table 2.2 and Figure 3.2). The third and most recent CPUE series for Japan shows a sharp threefold increase in the most recent years, 2006-2012. The single CPUE series for Taiwan for 2000-2010 fluctuated around an average value. The most recent Japanese CPUE during 1994-2012 and the Taiwanese CPUE during 2000-2012 were moderately positively correlated with a Pearson correlation coefficient of $\rho=0.50$.

## Bayesian Production Model

## Posterior Distribution Convergence

For the WCNPO swordfish stock, a plot of the autocorrelation function indicated a thinning interval of 25 which was large enough to address potential autocorrelation in the MCMC runs. Visual inspection of trace plots for the major parameters showed good mixing of the three MCMC chains (i.e., fully-sampling the parameter space), and also indicated convergence of the MCMC chains. For all parameters, the Gelman and Rubin statistic, including the variance terms, equaled 1, which indicated convergence of the MCMC chains. Similarly, the Heidelberger and Welch test did not reject the hypothesis that the MCMC chains were stationary at the $95 \%$ confidence level for any of the parameters. Overall, these diagnostics indicated that the posterior distribution of the model parameters was adequately sampled with the MCMC simulations.

For the EPO swordfish stock, all key model parameters (intrinsic growth rate, carrying capacity, production function shape parameter, and catchability coefficients) and biological reference points converged according to the Geweke diagnostic (Geweke 1992), Gelman and Rubin diagnostic (Gelman and Rubin 1992), and the Heidelberger and Welch stationarity and halfinterval tests. A visual inspection of model parameter posterior distribution density plots indicated that these densities were smooth and unimodal for all parameters as expected for a convergent sequence of MCMC samples. Overall, the convergence diagnostics that were examined indicated that the MCMC samples generated from the generalized production model had numerically converged to the posterior distribution.

## Production Model Fits to CPUE

For the WCNPO stock, the predicted CPUE indices were compared to the observed CPUE for each model to determine the adequacy of model fit (Figure 4.1). Plots of standardized residual diagnostics by fishery for the base-case model indicated a good fit to the long-term Japanese longline CPUE and more variable fits to the shorter CPUE time series (Figure 4.1). Fits of other candidate runs were also examined and were summarized in Chang et al. (2014) but are not
presented here. A summary table of residual patterns, normality test results, RMSE values and DIC values showed that several patterns were immediately apparent (Table 4.1 and Figure 4.1):

1) Models which included Taiwanese longline CPUE during 1969-1999 had residuals which showed non-random temporal patterns for both the two Taiwanese longline CPUE indices (19691999 and 2000-2012) and the Hawaii longline CPUE. The Taiwanese longline CPUE during 2000-2012 also failed the Shapiro-Wilk normality test ( $\mathrm{W}=0.82$, $\mathrm{P}<0.05$ ).
2) The model fit using the alternative Japanese longline CPUE that did not include data from areas 8 and 9 showed a poorer fit $(\operatorname{RMSE}=2.716)$ and appeared to have a non-random residual pattern for the Hawaii CPUE in comparison to model fits which included Japanese data from areas 8 and $9(\mathrm{RMSE}=2.273)$.
3) Assuming a higher CV for the prior distribution of P1 (CV=50\%) did not produce an overall improvement to model fit to the CPUE indices and also had a poorer fit and residual pattern for the Hawaii CPUE series ( $\mathrm{RMSE}=2.758$ ).
4) DIC values were compared among models with the same data structure. Results indicated that the minimum value of DIC ( $\mathrm{DIC}=-185.49$ ) was achieved by model selected as the base case model. DIC values for the other two viable candidate models were 8.48 and 2.90 units higher than for the base case model (Chang et al. 2014), respectively. Based on all this information, the base case model for stock status determination was agreed upon by the ISC Billfish Working Group.

For the base case model fit to Japanese CPUE, predicted CPUE values fluctuated about the observed CPUE time series and the standardized residuals had no time trend and were normally distributed (Figure 4.1). However, the Taiwanese longline CPUE fit had a pattern of consecutive negative residuals in the late-2000s and the standardized residuals failed the normality test at significance level of $0.05(\mathrm{~W}=0.81, \mathrm{P}=0.001)$. Fits to the Hawaii longline CPUE appeared to have no trend in residuals and the standardized residuals were normally distributed. Overall, the base case model fits to the WCNPO Pacific swordfish CPUE indicated that there was a good fit to the Japanese longline CPUE and a minor lack of fit to the Taiwanese longline CPUE.

For the EPO stock, the base case model fit the standardized CPUE series adequately (Figure 4.2). Standardized residuals for the first and second Japanese CPUE series appeared to be random and predicted CPUE fluctuated randomly about the observed CPUE. Standardized residuals for the third Japanese CPUE series did not appear random and the predicted CPUE was an underestimate of the observed Japanese CPUE during 2006-2012 when the CPUE series exhibited a twofold increase (Figure 4.2). During 2006-2012, the residuals for the Taiwanese CPUE also did not appear to have a flat trend although the magnitude of those residuals was relatively small. However, the Shapiro-Wilks normality test indicated that standardized residuals from each CPUE series were normally distributed ( $\mathrm{P}>0.05$ ). Standardized residuals of the first Japanese CPUE, third Japanese CPUE, and Taiwanese CPUE exhibited some time trend ( $\mathrm{P} \leq 0.01$ ) according to a regression of standardized residuals against time. Bartlett's test showed that the variances of standardized residuals for the first and third Japanese CPUE were not homogeneous ( $\mathrm{P}<0.05$ ), but variance was homogeneous for the Taiwanese CPUE ( $\mathrm{P}>0.05$ ). Standardized residuals for the second Japanese CPUE series showed no time trend ( $\mathrm{P}>0.05$ ) and had homogeneous variance ( $\mathrm{P}>0.05$ ). Overall, the CPUE fits were judged to be adequate, albeit variable.

Based on RMSE values, model fits were best for the first two Japanese CPUE series (1952-1974, 1975-1993), followed by the Taiwanese series (2000-2012) (Table 4.2). The poorest fit (highest RMSE value) came from the model fit to the third Japanese CPUE series from 1994-2012, which reflected the difficulty of fitting the model to the high values of CPUE in recent years. The correlation coefficients for all Japanese CPUE series were greater than one-half ( $\rho>0.50$ ) and indicated a generally good model fit to CPUE trend. The correlation coefficient for the Taiwan CPUE series was a moderately good fit to CPUE trend with a value of $\rho=0.43$. Overall, the base case model fit was judged to provide an adequate fit to the CPUE series with some exceptions for the most recent years of high observed CPUE for the Japanese longline fishery.

## Estimated Parameters, Quantities of Interest, and Reference Points

For the WCNPO stock, estimates of the mean and standard deviation of model parameters, quantities of interest, and MSY-based reference points were tabulated (Table 5.1). Estimates of posterior densities of the parameters $\mathrm{r}, \mathrm{K}, \mathrm{M}, \sigma^{2}, \tau^{2}$, and P1 were smooth and unimodal. Summaries of posterior quantiles of parameters and quantities of interest were provided and showed that the marginal posteriors generally were right-skewed. The marginal posterior for r had a median of $\mathrm{r}=0.54$ ( $0.28-1.1195 \%$ C.I.), similar to the prior mean. Although both the posterior and prior for K had a peak around 120,000 metric tons, the posterior was much less dispersed than the prior. The marginal posteriors for M and $\mathrm{P}_{1}$ had median values of 0.89 (0.36$2.0895 \%$ C.I.) and 0.84 ( $0.69-1.0395 \%$ C.I.), respectively, and these values were slightly different from the prior means. Although diffuse priors were assigned to the process error and observation error variances, the posterior error variances were less dispersed than the priors, which indicated the data reduced uncertainty for the error variances. Furthermore, the observation error variances were greater than the process error variance. The marginal posteriors for MSY, HMSY, and BMSY were slightly right-skewed and were centered at median values of MSY $=14,730$ metric tons, $\mathrm{HMSY}=0.25$, and $\mathrm{BMSY}=59,520$ metric tons, respectively.

Parameter estimates from this 2014 update were generally similar to those from the 2009 assessment (Table 5.1). The estimate of $K$ scaled with exploitable biomass and was slightly higher in the current assessment (median $121,000 \mathrm{mt}$ ) compared to the 2009 assessment (113,000 mt ). The estimates of r and M in the current assessment also did not differ substantially from the estimated values in 2009. As a result, estimates of exploitable biomass to maximize surplus production, BMSY, and the maximum surplus production, MSY, from the 2014 assessment were all slightly higher than those values from the 2009 assessment.

Exploitable biomass of WCNPO swordfish fluctuated at or above BMSY throughout the assessment time horizon (Table 6.1 and Figure 5.1). As expected, there was an inverse pattern between estimated biomass and harvest rate as harvest rate fluctuated at or below HMSY. Trends in exploitable biomass and harvest rate from the current assessment are very similar to those from the 2009 assessment (Figure 5.1). After several years of high catches, harvest rates increased to fluctuate around HMSY during 1956-1961. As a result, exploitable biomass decreased to $69,000 \mathrm{mt}$ in 1962, and then fluctuated around $70,000 \mathrm{mt}$ for a decade. Harvest rates fluctuated around $50 \%$ of HMSY from the mid-1960s to the late-1980s. Concurrently, exploitable biomass increased to a peak of $121,000 \mathrm{mt}$ in 1987, or roughly two-fold higher than BMSY. Due to increased swordfish catches during the 1990s, harvest rates increased to fluctuate about HMSY and the exploitable biomass gradually declined to roughly BMSY in 1996. The

WCNPO swordfish catch has had a declining trend since 2007, with harvest rates fluctuating around $50 \%$ of HMSY (Figures 2.1 and 5.1). For the recent 10 years, exploitable biomass has been relatively stable and fluctuated above BMSY (around 70,000 mt). The probabilities of exploitable biomass being below BMSY and harvest rate exceeding HMSY in 2012 were estimated to be 0.14 and 0 , respectively. The Kobe plot showed that the WCNPO swordfish stock does not appear to have been depleted or to have experienced overfishing throughout most of the assessment time horizon of 1951-2012 (Figure 6.1).

For the EPO stock, estimates of the mean and standard deviation of model parameters, quantities of interest, and MSY-based reference points were tabulated (Table 5.2). The intrinsic growth rate was estimated to be $r=0.46$ and the carrying capacity was $K=65,000 \mathrm{mt}$ with an maximum sustainable yield of MSY=5,490 mt. Biomass to produce maximum sustainable yield was estimated to be about $50 \%$ of the carrying capacity at BMSY $=31,200 \mathrm{mt}$, and harvest rate to produce maximum sustainable yield was estimated to be HMSY $=0.18$. The initial proportion of biomass to carrying capacity was $\mathrm{P}_{1}=0.88$, close to the initially assumed value of 0.90 . The production shape parameter was estimated as $\mathrm{M}=0.93$, close to a symmetric Schaefer curve, but was imprecisely estimates with a standard deviation of 0.71 ( $\mathrm{CV}=76 \%$ ).

Time series of model estimates of exploitable biomass and harvest rate over the assessment time horizon were also summarized (Table 6.2 and Figure 5.2). Exploitable biomass was initially near carrying capacity in the early-1950s and since then, has fluctuated and ranged from 30,000 to $60,000 \mathrm{mt}$, generally remaining above $B_{M S Y}$ throughout the assessment time horizon. It is notable that the estimated $95 \%$ confidence intervals for exploitable biomass are wide and large enough that the lower $95 \%$ confidence limit falls below $B_{M S Y}$ over much of the time period (Figure 5.2). Harvest rates were initially low and steadily increased through time, and likely exceeded $H_{M S Y}$ in 1998, 2002, 2003, and also the most recent year, 2012. Trends of exploitable biomass and harvest rate in relation to MSY-based biological reference points were presented as a Kobe plot (Figure 6.2). This plot illustrated that overfishing has likely occurred in a few years in the history of the fishery, and is likely to be occurring in some recent years. In 2012, swordfish in the EPO were experiencing overfishing with a $55 \%$ probability.

## Sensitivity Analyses

For the WCNPO swordfish stock, the sensitivity analyses for the input prior means of the four parameters showed that the model results were robust to changes in the prior assumptions (Table 7.1 and Figures 7.1.1 and 7.1.2). The trends of relative biomass B/BMSY were almost the same except for the model runs with high or low P1 prior mean (Table 7.1 and Figures 7.1.1 and 7.1.2). However, the status of relative biomass in 1997 and 1998 appeared to depend on the sensitivity scenario. The impact of using lower $r$ and P1 prior means was to produce more pessimistic estimates of stock status. A similar inverse pattern was found in the results for relative harvest rate H/HMSY. Overall, the sensitivity analyses suggested that prior assumptions were not driving the results of the base case WCNPO swordfish production model.

For the EPO swordfish stock, the sensitivity analysis using high (+25\%) and low ( $-25 \%$ ) values of input prior means for the parameters $R, K, S$, and $P_{1}$ generally indicated that the model results were robust to changes in the prior assumptions (Table 7.2 and Figures 7.2.1 and 7.2.2). The trend and scale of exploitable biomass and harvest rate estimates were also robust to the high and
low alternative prior means (Table 7.2 and Figures 7.2.1 and 7.2.2). Overall, this suggested that the priors were not unduly influential for the base case EPO swordfish production model results.

## Retrospective Analyses

For the WCNPO stock, retrospective analyses showed that the time series of exploitable biomass estimates produced with the removal of annual assessment data in successive model runs matched very well with the base case model (Figure 8.1). For WCNPO swordfish the value of Mohn's (1999) DR statistic was DR=-0.06 for exploitable biomass, which supported the fact that there was no retrospective pattern for the estimates of exploitable biomass.

For the EPO stock, the retrospective analyses revealed a clear and consistent pattern. Terminal year estimates of biomass and harvest rate appeared to be biased with an underestimation of exploitable biomass and an overestimation of harvest rate (Figure 8.2). The cause of the retrospective pattern in the data has not yet been determined. Any management decisions based on the results of this assessment should consider the fact that there is a clear retrospective pattern in estimates of quantities of interest.

## Projections

For the WCNPO stock, projections results were summarized for each of the eight harvest scenarios (Table 8.1 and Figures 9.1.1 and 9.1.2). Stochastic projections indicated there exploitable biomass would likely remain above BMSY through 2016 under the status quo catch or status quo harvest rate scenarios (Table 8.1 and Figure 9.1.1). For the high harvest rate scenarios MaxF, $1.5 *$ FMSY, $1.25 *$ FMSY), exploitable biomass would be projected to decline below BMSY by 2016 (Table 8.1 and Figures 9.1.1 and 9.1.2). Projected harvest rates would exceed the MSY-based overfishing threshold for the high harvest rate scenarios, and in particular, the stock would not be experiencing overfishing during 2014-2016 under the status quo catch or harvest rate scenarios.

For the EPO stock, stochastic projections revealed that exploitable biomass will likely have a decreasing trajectory in the near future under all of the harvest scenarios examined (Table 8.2and Figures 9.2.1 and 9.2.2). Under the high harvest scenarios (status quo catch, MaxF, and $1.5 * F_{M S Y}$ ), exploitable biomasses were projected to decline to be roughly equal to $B_{M S Y}$ in 2016 (Table 8.2and Figures 9.2.1 and 9.2.2). These high relative harvest rate scenarios, including $1.25 * F_{M S Y}$, also resulted in harvest rates above $H_{M S Y}$ (Table 8.2). Under the status quo harvest rate scenario, exploitable biomass was projected to decline to about $40,000 \mathrm{mt}$ by 2016 , well above the BMSY level. The lowest future harvest scenario of $0.50 * F_{M S Y}$ resulted in the smallest decline of exploitable biomass, which was projected to decline from 50,300 mt in 2013 to 48,400 mt in 2016. Overall, the projections showed that if recent high catch levels persist, exploitable biomass will very likely decrease and a moderate risk of overfishing will likely continue to occur.

## Risk Analyses

For the WCNPO stock, the risk analyses showed that there was virtually no chance of the stock being overfished or experiencing overfishing in 2016 (Table 9.1 and Figure 10) if current catch levels of about $10,000 \mathrm{mt}$ are maintained. Catches would need to increase to average roughly $15,000 \mathrm{mt}$ to have a moderate ( $50 \%$ chance) risk of overfishing and would need to increase to about $25,000 \mathrm{mt}$ to have a moderate risk of the stock being overfished in 2016 (Figure 10).

For the EPO stock, the risk analyses showed that the probabilities of overfishing and becoming overfished increased as projected catch increased in the future (Table 9.2 and Figure 10). Maintaining the current EPO swordfish catch level would lead to a moderate risk of overfishing in 2016 but would lead to virtually no chance of the stock being overfished in 2016. In particular, future catch levels would need to be reduced to average approximately $9,700 \mathrm{mt}$, or slightly below the recent average catch, for the probability of overfishing to be below a moderate risk threshold of $50 \%$. In contrast, at the $9,700 \mathrm{mt}$ catch level there would a roughly $0 \%$ chance of becoming overfished by 2016 (Figure 10). In comparison, a reduced catch level of $5,800 \mathrm{mt}$, or slightly above MSY, would result in a reduced probability of overfishing of only about $12 \%$ in 2016 in comparison to the status quo harvest rate scenario. Overall, the EPO swordfish stock is perceived to be at a high abundance level above BMSY. However, any increases in catch to fish down the stock should be considered in light of the observed retrospective pattern of biomass estimation in recent years.

## DISCUSSION

For the WCNPO stock, the Bayesian estimation framework provided conservation information that accounted for uncertainty in estimates of stock status relative to biological reference points. This is important for effectively conveying stock assessment results to fisheries managers and stakeholders. The probabilistic interpretation of stock status showed that it was very unlikely that the WCNPO swordfish population biomass was below BMSY in 2012 ( $\operatorname{Pr(B2012~<~}$
BMSY) $=0.14$ ). Similarly, it was extremely unlikely that the swordfish population was being fished in excess of HMSY in $2012(\operatorname{Pr}(\mathrm{H} 2012>\mathrm{HMSY})<0.005)$.

For the EPO stock, the generalized production model produced estimates and associated uncertainty of parameters, biological reference points, stock status, and future stock status given different harvest scenarios. Results indicated that overfishing may be occurring in the swordfish longline fishery in the Eastern Pacific Ocean. This result reflects changes to the production model fits to the revised catch and CPUE time series, including the fit to the increased catches and increased Japanese CPUE in recent years. The high increase exhibited by Japanese CPUE in recent years, however, was not matched by an increase in the Taiwanese CPUE. There is a 55\% probability that overfishing was occurring in 2012, while there is a $0 \%$ chance that the stock was overfished in 2012. If the 2012 high catch levels persist, the moderate risk of overfishing will also persist. Therefore, while the exploitable biomass of the EPO swordfish stock is likely at a healthy level, the most recent catch levels of $10,000 \mathrm{mt}$, or roughly two times higher than the estimated MSY, are not likely to sustainable in the long term.

During the model selection process and prior to settling on the base case model for EPO swordfish, it was observed that model runs that included the early Taiwanese CPUE series (1968-1999) (Sun et al. 2014) failed to converge. As a result, the early Taiwanese CPUE series was excluded from the final base case model. Future assessments should explore the inclusion of this CPUE series using updated catch and CPUE data, and perhaps alternative standardization models.

We caution that our analysis revealed a clear retrospective pattern in the EPO assessment results, with underestimation of exploitable biomass and overestimation of harvest rate. Any
management decisions should take this pattern into account. We recommend that further assessment work on North Pacific swordfish should be conducted to determine whether the retrospective pattern can be accounted for, and further work should also investigate using more detailed biological data with age- or length-structured models.

Applying a Bayesian estimation framework allowed us to make clear statements about the degree of confidence in estimated quantities (Ellison 2004), including biological reference points and the effect of various future harvest scenarios on the stock. By providing probabilities of overfishing and becoming overfished for future harvest scenarios, it is hoped that this information would enable managers to implement a precautionary approach to swordfish fishery management in which acceptable risk levels for undesirable outcomes are selected and decision tables are applied to judge the efficacy of alternative management options (Hilborn and Peterman 1996, McAllister and Kirkwood 1998). A notable result from the use of the Bayesian estimation framework is the large $95 \%$ confidence intervals for biomass estimates indicating moderately high uncertainty over the time series of the fishery (1951-2012) and also in the projections and risk analyses.

Although a single stock has generally been assumed for assessment purposes, fisheries stock assessment scientists recognized that not all exploited species fit easily into a unit stock definition. In 2009, the ISC (2009) swordfish assessment indicated that the North Pacific swordfish population was be estimated to be a smaller- (lower K) and more productive stock (higher $r$ ) under the single-stock scenario than as a combination of two stocks under the twostock scenario. While an update the stock assessment for the swordfish in WCNPO area based on the two stocks scenario has been provided, we suggest that alternative swordfish stock structure hypotheses may need to be included in future assessments to address the uncertainty associated with stock structure.

Using a Bayesian estimation approach allowed us to make clear statements about the degree of confidence and uncertainty in estimated quantities. However, it is important to note that the choice of prior distributions can alter posterior estimates of stock status, especially when data quality is questionable (Booth and Quinn, 2006). Although the sensitivity analyses suggested that the prior mean were not driving the results of the base case, we suggest that it is important in future work to explore the robustness of our stock assessment models to different prior distribution functions (e.g., uniform). We also suggest the development and refinement of informative priors based on demographic analyses to reduce the estimation uncertainty (McAllister et al., 2001).

Swordfish are known to be sexually dimorphic. For example, swordfish females mature later than males and the sex-ratio varies with length (DeMartini et al., 2000). These phenomena have implications for fishery selectivity and hence fishing-induced mortality. Therefore, we also recommend that further assessment work on WCNPO swordfish consider more detailed biological data with sex-specific and age- or length-structured models to better approximate the population dynamics.

## ACKNOWLEDGMENTS

We sincerely thank the member countries of the ISC, the IATTC, and the WCPFC for their help in preparing and providing information for this North Pacific swordfish stock assessment update.

## REFERENCES

Alvarado Bremer, J.R., Hinton, M.G., Greig, T.W., 2006. Evidence of spatial genetic heterogeneity in Pacific swordfish (Xiphias gladius L.) revealed by the analysis of ldh-A sequences. Bull. Mar. Sci. 79, 493-503.

Best, N.G., Cowles, M.K., Vines, S.K., 1995. CODA Manual Version 0.30. MRC, Biostatistics Unit, Cambridge, pp. 41.

Booth, A., Quinn, T.J.I., 2006. Maximum likelihood and Bayesian approaches to stock assessment when data are questionable. Fish. Res. 80, 169-181.

Brodziak, J. 2007. An investigation of alternative production models to assess the Hawaiian bottomfish complex. Administrative Report H-07-01, Pacific Islands Fisheries Science Center, National Marine Fisheries Service, NOAA, Honolulu, HI, 96822.

Brodziak, J. 2009. Potential natural mortality rates of North Pacific swordfish. International Scientific Committee for Tuna and Tuna-Like Species in the North Pacific, Billfish Working Group. ISC/09/BILLWG-1/13.

Brodziak, J. and G. Ishimura. 2009. Development of Bayesian surplus production models for assessing the North Pacific swordfish population. International Scientific Committee for Tuna and Tuna-Like Species in the North Pacific, Billfish Working Group. ISC/09/BILLWG-2/02.

Brodziak, J. 2010. Update of the production model assessment of the Eastern Pacific swordfish stock (Xiphias gladius) in 2010. International Scientific Committee for Tuna and Tuna-Like Species in the North Pacific, Billfish Working Group. ISC/10/BILLWG-1/02.

Chang, Y.-J., A. Yau, and J. Brodziak. 2014. Stock assessment of Western and Central North Pacific Ocean swordfish (Xiphias gladius) through 2012. International Scientific Committee for Tuna and Tuna-Like Species in the North Pacific, Billfish Working Group. ISC/14/BILLWG1/02.

Congdon, P. 2001. Bayesian statistical modeling. Wiley, New York.
DeMartini, E.E., Uchiyama, J.H., Williams, H.A., 2000. Sexual maturity, sex ratio, and size composition of swordfish, Xiphias gladius, caught by the Hawaii-based pelagic longline fishery. Fish. Bull. 98, 489-506.

DeMartini, E. E., J. H. Uchiyama, R. L. Humphreys Jr, J. D. Sampaga, and H. A. Williams. 2007. Age and growth of swordfish (Xiphias gladius) caught by the Hawaii-based pelagic longline fishery. Fishery Bulletin 105:356-367.

DeMartini, E. E., J. H. Uchiyama, and H. A. Williams. 2000. Sexual maturity, sex ratio, and size composition of swordfish, Xiphias gladius, caught by the Hawaii-based pelagic longline fishery. Fishery Bulletin 98:489-506.

Dreyfus, M., L. A. Fleischer, J. L. Castillo-Géniz, L. V. González-Ania, A. Liedo-Galindo, J. Tovar-Ávila, P. A. U. Ramírez, and J. G. D. Uribe. 2013. National report of Mexico. International Scientific Committee for Tuna and Tuna-Like Species in the North Pacific. ISC/13/PLENARY/08.

Ellison, A. M. 2004. Bayesian inference in ecology. Ecology letters 7:509-520.
Gelman, A. and D. B. Rubin. 1992. Inference from iterative simulation using multiple sequences. Statistical science 7:457-472.

Geweke, J. 1992. Evaluating the accuracy of sampling-based approaches to the calculation of posterior moments. Pages 169-193 Bayesian Statistics 4. Oxford University Press, Oxford, U.K. . Gilks, W. R., S. Richardson, and D. J. Spiegelhalter. 1996. Markov chain Monte Carlo in practice. CRC press, London.

Heidelberger, P. and P. D. Welch. 1983. Simulation run length control in the presence of an initial transient. Operations Research 31:1109-1144.

Hilborn, R., Walters, C.J., 1992. Quantitative fisheries stock assessment: choice, dynamics and uncertainty. Chapman and Hall, New York, pp. 570.

Hilborn, R. and R. Peterman. 1996. The development of scientific advice with incomplete information in the context of the precautionary approach. Pages 77-97 FAO Fisheries Technical Paper. FAO, Lysekil, Sweden.

Hinton, M. G. and M. N. Maunder. 2011. Status of Swordfish in the Eastern Pacific Ocean in 2010 and Outlook for the Future. Inter-American Tropical Tuna Commission, Scientific Advisory Committee. SAC-02-09.

Ichinokawa, G. and J. Brodziak. 2008. Stock boundary between possible swordfish stocks in the northwest and southwest Pacific judged from fisheries data of Japanese longliners. International Scientific Committee for Tuna and Tuna-Like Species in the North Pacific, Billfish Working Group. ISC/08/BILLWG-SS/04.

ISC. 2009. ISC Plenary, Annex 7. Report of the Billfish Working Group Workshop 19-26 May, 2009, Busan, Korea., International Scientific Committee for Tuna and Tuna-Like Species in the North Pacific, Billfish Working Group.

Ito, R., Childers, J.J., 2014. U.S. Swordfish Fisheries in the North Pacific Ocean. ISC/14/BILLWG-1/06.

Kimoto, A., M. Kanaiwa, and K. Yokawa. 2014. Update of the catch per unit effort (CPUE) distribution of swordfish (Xiphias gladius) by the Japanese offshore and distantwater longline fishery in the Pacific. International Scientific Committee for Tuna and Tuna-Like Species in the North Pacific, Billfish Working Group. ISC/14/BILLWG-1/07.

Kimoto, A. and K. Yokawa. 2014. Updated catch amount of swordfish (Xiphias gladius) by the Japanese coastal, offshore, and distant-water longline fishery in the Pacific. International Scientific Committee for Tuna and Tuna-Like Species in the North Pacific, Billfish Working Group. ISC/14/BILLWG-1/04.

Kleiber, P. and K. Yokawa. 2004. MULTIFAN-CL assessment of swordfish in the North Pacific. International Scientific Committee for Tuna and Tuna-Like Species in the North Pacific, Swordfish Working Group. ISC/04/SWO-WG/07.

Lunn, D., Thomas, A., Best, N., Spiegelhalter, D., 2000. WinBUGS: a Bayesian modelling framework: concepts, structure, and extensibility. Stat. Comput. 10, 325-337.

McAllister, M., E. Babcock, E. K. Pikitch, and M. H. Prager. 2000. Application of a nonequilibrium generalized production model to South and North Atlantic swordfish: Combining Bayesian and demographic methods for parameter estimation. Col. Vol. Sci. Pap. ICCAT 51:1523-1550.

McAllister, M. and G. Kirkwood. 1998. Bayesian stock assessment: a review and example application using the logistic model. ICES Journal of Marine Science 55:1031-1060.

McAllister, M., E. K. Pikitch, and E. Babcock. 2001. Using demographic methods to construct Bayesian priors for the intrinsic rate of increase in the Schaefer model and implications for stock rebuilding. Canadian Journal of Fisheries and Aquatic Sciences 58:1871-1890.

Meyer, R. and R. B. Millar. 1999. BUGS in Bayesian stock assessments. Canadian Journal of Fisheries and Aquatic Sciences 56:1078-1087.
Mohn, R. 1999. The retrospective problem in sequential population analysis: An investigation using cod fishery and simulated data. ICES Journal of Marine Science: Journal du Conseil 56:473-488.

Plummer, M., N. Best, K. Cowles, and K. Vines. 2006. CODA: Convergence diagnosis and output analysis for MCMC. R news 6:7-11.

R Development Core Team. 2013. R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria.

SERNAPESCA. 2007-2012. Anuario Estadistico de Pesca y Acuicultura. Servicio Nacional de Pesca y Acuicultura, Valparaiso, Chile.

Spiegelhalter, D.J., Thomas, A., Best, N.G., Carlin, B.P., vander Linde, A., 2002. Bayesian measures of model complexity and fit. J. R. Stat. Soc. B 64, 583-640.

Spiegelhalter, D., A. Thomas, N. Best, and D. Lunn. 2003. WinBUGS user manual.
Sturtz, S., Ligges, U., Gelman, A., 2005. R2WinBUGS: A package for running WinBUGS from R. J. Stat. Soft. 12.

Sun, C. L., N. J. Su, and S. Z. Yeh. 2014. Standardized CPUE of swordfish (Xiphias gladius) for the Taiwanese distant-water tuna longline fishery, based on a two stock scenario in the North Pacific. . International Scientific Committee for Tuna and Tuna-Like Species in the North Pacific, Billfish Working Group. ISC/14/BILLWG-1/07.

Tagami, D., Wang, H., Chang, Y.J., 2014. Spatial distribution of swordfish catches for longline fisheries in the Western and Central North Pacific and Eastern Ocean. ISC/14/BILLWG-1/03.

Uchiyama, J.H., Humphreys, R.L., 2007. Revised review table of vital rates and life history parameters for striped marlin, swordfish, and blue marlin in the North Pacific Ocean (February 2007). Pacific Islands Fisheries Science Center, Honolulu, HI Unpublished Pers. Comm.

Walsh, W., Brodziak, J., 2014. Catch rate standardization for swordfish Xiphias gladius in the shallow-set sector of the Hawaii-based pelagic longline fishery: 1995-2012. ISC/14/BILLWG1/05.

Wang, S.-P., C.-L. Sun, A. E. Punt, and S.-Z. Yeh. 2005. Evaluation of a sex-specific agestructured assessment method for the swordfish, Xiphias gladius, in the North Pacific Ocean. Fisheries Research 73:79-97.

Wang, S.-P., C.-L. Sun, A. E. Punt, and S.-Z. Yeh. 2007. Application of the sex-specific agestructured assessment method for swordfish, Xiphias gladius, in the North Pacific Ocean. Fisheries Research 84:282-300.

Ward, P. and S. Elscot. 2000. Broadbill swordfish: Status of world fisheries. Bureau of Rural Sciences, Commonwealth Department of Agriculture, Fisheries and Forestry, Canberra, Australia.

Table 1.1. Swordfish (Xiphias gladius) catches (metric tons) in the Western and Central Pacific Ocean by fisheries, 1951-2012. A "-" indicates no effort or data not available, and "0" indicates less than 1 metric ton.

| Year | Japan |  | Taiwan |  | USA |  | Korea | ${ }^{\dagger}$ Other | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{gathered} \text { Deepwater } \\ \& \\ \text { offshore } \\ \text { longline } \\ \hline \end{gathered}$ | Other | Deepwater longline | Offshore <br> longline <br> \& other | Longline <br> (Hawaii) | Other | Longline |  |  |
| 1951 | 7245 | 4432 | - | - | - | - | - | - | 11677 |
| 1952 | 8888 | 2801 | - | - | - | - | - | - | 11689 |
| 1953 | 10794 | 1612 | - | - | - | - | - | - | 12405 |
| 1954 | 12543 | 1047 | - | - | - | - | - | - | 13591 |
| 1955 | 13050 | 1047 | - | - | - | - | - | - | 14097 |
| 1956 | 14590 | 890 | - | - | - | - | - | - | 15480 |
| 1957 | 14207 | 983 | - | - | - | - | - | - | 15190 |
| 1958 | 18510 | 1209 | - | - | - | - | - | - | 19719 |
| 1959 | 17181 | 1031 | - | 518 | - | - | - | - | 18731 |
| 1960 | 19983 | 1342 | - | 647 | - | - | - | - | 21972 |
| 1961 | 19398 | 1432 | - | 391 | - | - | - | - | 21221 |
| 1962 | 9950 | 1508 | - | 556 | - | - | - | - | 12014 |
| 1963 | 9644 | 922 | - | 361 | - | - | - | - | 10926 |
| 1964 | 5594 | 1183 | 0 | 368 | - | - | - | - | 7145 |
| 1965 | 7506 | 2249 | 0 | 358 | - | - | - | - | 10113 |
| 1966 | 8809 | 1897 | 0 | 520 | - | - | - | - | 11226 |
| 1967 | 9845 | 1125 | 0 | 681 | - | - | - | - | 11651 |
| 1968 | 8067 | 1839 | 0 | 775 | - | - | - | - | 10681 |
| 1969 | 7508 | 1920 | 0 | 850 | - | - | - | - | 10278 |
| 1970 | 5280 | 2223 | 0 | 909 | 5 | 622 | - | - | 9039 |
| 1971 | 5437 | 909 | 0 | 995 | 1 | 102 | 0 | - | 7444 |
| 1972 | 4814 | 891 | 0 | 873 | 0 | 175 | 0 | - | 6753 |
| 1973 | 4833 | 1307 | 0 | 979 | 0 | 403 | 0 | - | 7522 |
| 1974 | 4791 | 2193 | 0 | 1016 | 0 | 428 | 0 | - | 8428 |
| 1975 | 5835 | 3575 | 11 | 1052 | 0 | 570 | 0 | - | 11043 |
| 1976 | 6386 | 4747 | 10 | 807 | 0 | 55 | 0 | - | 12005 |
| 1977 | 7452 | 3505 | 3 | 683 | 17 | 337 | 165 | - | 12162 |
| 1978 | 7532 | 3769 | 0 | 558 | 9 | 1712 | 53 | - | 13633 |
| 1979 | 8168 | 2246 | 7 | 694 | 7 | 386 | - | - | 11508 |
| 1980 | 5655 | 3038 | 11 | 679 | 5 | 788 | 47 | - | 10223 |
| 1981 | 6638 | 2774 | 1 | 681 | 3 | 746 | - | - | 10843 |
| 1982 | 5312 | 2392 | 1 | 904 | 5 | 1111 | 39 | - | 9764 |
| 1983 | 7318 | 2239 | 0 | 949 | 5 | 1758 | 9 | - | 12278 |
| 1984 | 7001 | 2458 | 0 | 997 | 3 | 2838 | 42 | - | 13339 |
| 1985 | 9114 | 2402 | 0 | 825 | 2 | 3399 | 22 | - | 15764 |
| 1986 | 8160 | 2480 | 0 | 667 | 2 | 2469 | 7 | - | 13785 |


| 1987 | 8695 | 2054 | 1 | 1518 | 24 | 1795 | 35 | - | 14122 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 8144 | 2112 | 0 | 1040 | 24 | 1638 | 21 | - | 12979 |
| 1989 | 5942 | 2741 | 4 | 1529 | 218 | 1361 | 30 | - | 11825 |
| 1990 | 5390 | 1909 | 5 | 1463 | 2436 | 1238 | 41 | - | 12482 |
| 1991 | 4377 | 1483 | 10 | 1570 | 4508 | 1035 | 3 | - | 12986 |
| 1992 | 6911 | 2471 | 2 | 1716 | 5700 | 1540 | 5 | - | 18345 |
| 1993 | 7955 | 2043 | 58 | 1484 | 5909 | 1768 | 11 | - | 19228 |
| 1994 | 7015 | 2127 | 0 | 1374 | 3176 | 1604 | 49 | - | 15345 |
| 1995 | 6005 | 2412 | 71 | 1360 | 2713 | 1165 | 7 | - | 13733 |
| 1996 | 6260 | 2141 | 10 | 733 | 2502 | 1203 | 11 | - | 12860 |
| 1997 | 6250 | 1992 | 20 | 1419 | 2881 | 1315 | 69 | - | 13946 |
| 1998 | 5590 | 2207 | 22 | 1219 | 3263 | 1416 | 100 | - | 13817 |
| 1999 | 5292 | 2241 | 63 | 1446 | 3100 | 1943 | 102 | - | 14187 |
| 2000 | 5398 | 2480 | 64 | 3476 | 2949 | 2630 | 147 | - | 17144 |
| 2001 | 5194 | 1915 | 121 | 3903 | 220 | 2181 | 255 | - | 13789 |
| 2002 | 5199 | 2370 | 155 | 3793 | 204 | 1715 | 284 | - | 13720 |
| 2003 | 4794 | 2442 | 144 | 3554 | 147 | 2156 | 247 | - | 13484 |
| 2004 | 4939 | 2834 | 502 | 3327 | 213 | 1200 | 300 | - | 13315 |
| 2005 | 5054 | 2777 | 269 | 3505 | 1622 | 307 | 339 | 297 | 14170 |
| 2006 | 5805 | 2897 | 203 | 3891 | 1211 | 523 | 389 | 133 | 15051 |
| 2007 | 5916 | 3337 | 191 | 3744 | 1735 | 555 | 170 | 151 | 15799 |
| 2008 | 3979 | 2960 | 162 | 3443 | 2014 | 478 | 351 | 244 | 13631 |
| 2009 | 3729 | 2710 | 147 | 3222 | 1817 | 306 | 280 | 163 | 12375 |
| 2010 | 3660 | 1918 | 231 | 2324 | 1676 | 119 | 278 | 463 | 10670 |
| 2011 | 2430 | 1320 | 366 | 2999 | 1623 | 237 | 256 | 226 | 9456 |
| 2012 | 2446 | 1680 | 576 | 3049 | 1418 | 110 | 245 | 338 | 9863 |

$\dagger$ catch data from Belize, Cook Islands, China, Spain, Fiji, Federated States of Micronesia, Kiribati, Marshall Islands, Papua New Guinea, Senegal, Tuvalu, and Vanuatu

Table 1.2. Swordfish (Xiphias gladius) catches (metric tons) in the Eastern Pacific Ocean by fisheries, 1951-2012. A ‘-‘ indicates no effort or data is not available, and " 0 " indicates catch of less than 1 metric ton. Japanese catch in 2011 and 2012 is provisional.

| Year | Japan | Taiwan |  | Spain | Mexico | Peru | Korea | China | Chile | ${ }^{\dagger}$ Other |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Offshore \& distant water longline | Distant water longline | Offshore longline | Longline | All gears | Longline | Longline |  |  |  | TOTAL |
| 1951 | 1 | - | - | 0 | - | - | 0 | - | - | - | 1 |
| 1952 | 1 | - | - | 0 | - | - | 0 | - | - | - | 1 |
| 1953 | 3 | - | - | 0 | - | - | 0 | - | - | - | 3 |
| 1954 | 20 | - | - | 0 | - | 700 | 0 | - | - | - | 720 |
| 1955 | 14 | - | - | 0 | - | 400 | 0 | - | - | - | 414 |
| 1956 | 9 | - | - | 0 | - | 600 | 0 | - | - | - | 609 |
| 1957 | 124 | - | - | 0 | - | 600 | 0 | - | - | - | 724 |
| 1958 | 80 | - | - | 0 | - | 400 | 0 | - | - | - | 480 |
| 1959 | 81 | - | - | 0 | - | 400 | 0 | - | - | - | 481 |
| 1960 | 118 | - | - | 0 | - | 400 | 0 | - | - | - | 518 |
| 1961 | 527 | - | - | 0 | - | 300 | 0 | - | - | - | 827 |
| 1962 | 961 | - | - | 0 | - | 400 | 0 | - | - | - | 1361 |
| 1963 | 1592 | - | - | 0 | - | 200 | 0 | - | - | - | 1792 |
| 1964 | 3066 | 0 | - | 0 | - | 900 | 0 | - | - | - | 3966 |
| 1965 | 1718 | 0 | - | 0 | - | 300 | 0 | - | - | - | 2018 |
| 1966 | 2029 | 0 | - | 0 | - | 200 | 0 | - | - | - | 2229 |
| 1967 | 1523 | 21 | - | 0 | - | 1300 | 0 | - | - | - | 2844 |
| 1968 | 2350 | 15 | - | 0 | - | 800 | 0 | - | - | - | 3165 |
| 1969 | 5944 | 6 | - | 0 | - | 1200 | 0 | - | - | - | 7150 |
| 1970 | 3995 | 24 | - | 0 | - | 2396 | 0 | - | - | - | 6415 |
| 1971 | 2118 | 14 | - | 0 | - | 185 | 0 | - | - | - | 2317 |
| 1972 | 2653 | 22 | - | 0 | 2 | 550 | 0 | - | - | - | 3227 |
| 1973 | 3491 | 19 | - | 0 | 4 | 1941 | 0 | - | - | - | 5455 |
| 1974 | 1869 | 22 | - | 0 | 6 | 470 | 0 | - | - | - | 2367 |
| 1975 | 2037 | 8 | - | 0 | - | 158 | 9 | - | - | - | 2212 |
| 1976 | 2951 | 31 | - | 0 | - | 295 | 29 | - | - | - | 3306 |
| 1977 | 2573 | 27 | - | 0 | - | 420 | 33 | - | - | - | 3053 |
| 1978 | 2149 | 6 | - | 0 | - | 436 | 35 | - | - | - | 2626 |
| 1979 | 1674 | 16 | - | 0 | 7 | 188 | 18 | - | - | - | 1903 |
| 1980 | 2131 | 7 | - | 0 | 380 | 216 | 62 | - | - | - | 2796 |
| 1981 | 1926 | 25 | - | 0 | 1575 | 91 | 153 | - | - | - | 3770 |
| 1982 | 1806 | 14 | - | 0 | 1365 | 154 | 97 | - | - | - | 3436 |
| 1983 | 1752 | 5 | - | 0 | 120 | 238 | 65 | - | - | - | 2180 |
| 1984 | 1039 | 9 | - | 0 | 47 | 343 | 65 | - | - | - | 1503 |
| 1985 | 1039 | 8 | - | 0 | 18 | 55 | 91 | - | - | - | 1211 |


| 1986 | 2054 | 11 | - | 0 | 422 | 21 | 198 | - | - | - | 2706 |
| :--- | :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 | 2683 | 25 | - | 0 | 550 | 73 | 334 | - | - | - | 3665 |
| 1988 | 2670 | 23 | - | 0 | 613 | 54 | 163 | - | - | - | 3523 |
| 1989 | 2158 | 103 | - | 0 | 690 | 3 | 151 | - | - | - | 3105 |
| 1990 | 2645 | 29 | - | 0 | 2650 | 1 | 645 | - | - | - | 5970 |
| 1991 | 2739 | 44 | - | 0 | 861 | 3 | 696 | - | - | - | 4343 |
| 1992 | 3676 | 16 | - | 0 | 1160 | 16 | 372 | - | - | - | 5240 |
| 1993 | 2696 | 13 | - | 0 | 812 | 76 | 385 | - | - | - | 3982 |
| 1994 | 2507 | 18 | - | 0 | 581 | 310 | 344 | - | - | - | 3760 |
| 1995 | 2140 | 2 | - | 0 | 437 | 7 | 399 | - | - | - | 2985 |
| 1996 | 2116 | 24 | - | 0 | 439 | 1013 | 568 | - | - | - | 4160 |
| 1997 | 2755 | 26 | - | 6 | 2365 | 24 | 707 | - | - | - | 5884 |
| 1998 | 2949 | 80 | - | 115 | 3603 | 98 | 675 | - | - | - | 7520 |
| 1999 | 1551 | 69 | - | 29 | 1136 | 15 | 561 | - | - | - | 3361 |
| 2000 | 2001 | 283 | - | 831 | 2216 | 2 | 817 | - | - | - | 6150 |
| 2001 | 3735 | 2095 | - | 245 | 780 | 2 | 517 | - | - | - | 7374 |
| 2002 | 2824 | 3088 | - | 303 | 465 | 14 | 391 | - | - | - | 7085 |
| 2003 | 2615 | 1648 | 72 | 534 | 671 | 26 | 182 | - | - | - | 5748 |
| 2004 | 1809 | 1375 | 54 | 1292 | 270 | 19 | 1060 | - | - | - | 5878 |
| 2005 | 1408 | 713 | 93 | 717 | 235 | 28 | 287 | - | - | - | 3480 |
| 2006 | 1297 | 915 | 114 | 366 | 347 | 63 | 132 | - | - | - | 3235 |
| 2007 | 1386 | 783 | 36 | 661 | 172 | 46 | 284 | 50 | 246 | 38 | 3701 |
| 2008 | 1634 | 427 | 12 | 390 | 242 | 124 | 424 | 660 | 312 | 37 | 4262 |
| 2009 | 2079 | 663 | 76 | 2546 | 394 | 25 | 687 | 573 | 391 | 38 | 7473 |
| 2010 | 2653 | 994 | 107 | 3780 | 222 | 5 | 398 | 858 | 472 | 143 | 9631 |
| 2011 | 3094 | 790 | 286 | 2364 | 257 | 50 | 715 | 1571 | 182 | 278 | 9586 |
| 2012 | 2986 | 815 | 694 | 2377 | 257 | 50 | 601 | 1552 | 221 | 357 | 9910 |

$\dagger$ catch data from Belize, French Polynesia, United States, and Vanuatu

Table 2.1. Catch per unit effort (CPUE, \# of swordfish/1000 hooks) used for assessment of swordfish (Xiphias gladius) in the Western and Central North Pacific Ocean. A '-' indicates no effort or data available. Calculated relative CVs are reported here, but model runs assumed a relative CV of 1 for all values of all CPUE series.

| Year | Japan <br> Deepwater \& offshore longline |  | Taiwan <br> Deepwater longline |  | Hawaii <br> Longline |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | CPUE | Relative CV | CPUE | Relative CV | CPUE | Relative CV |
| 1951 | - | - | - | - | - | - |
| 1952 | 0.20 | 1.79 | - | - | - | - |
| 1953 | 0.17 | 1.78 | - | - | - | - |
| 1954 | 0.24 | 1.78 | - | - | - | - |
| 1955 | 0.21 | 1.76 | - | - | - | - |
| 1956 | 0.17 | 1.75 | - | - | - | - |
| 1957 | 0.18 | 1.75 | - | - | - | - |
| 1958 | 0.25 | 1.75 | - | - | - | - |
| 1959 | 0.19 | 1.74 | - | - | - | - |
| 1960 | 0.21 | 1.74 | - | - | - | - |
| 1961 | 0.20 | 1.74 | - | - | - | - |
| 1962 | 0.19 | 1.73 | - | - | - | - |
| 1963 | 0.22 | 1.73 | - | - | - | - |
| 1964 | 0.20 | 1.73 | - | - | - | - |
| 1965 | 0.22 | 1.72 | - | - | - | - |
| 1966 | 0.22 | 1.72 | - | - | - | - |
| 1967 | 0.19 | 1.71 | - | - | - | - |
| 1968 | 0.16 | 1.72 | - | - | - | - |
| 1969 | 0.18 | 1.72 | - | - | - | - |
| 1970 | 0.19 | 1.71 | - | - | - | - |
| 1971 | 0.19 | 1.72 | - | - | - | - |
| 1972 | 0.18 | 1.73 | - | - | - | - |
| 1973 | 0.21 | 1.73 | - | - | - | - |
| 1974 | 0.24 | 1.72 | - | - | - | - |
| 1975 | 0.21 | 1.05 | - | - | - | - |
| 1976 | 0.24 | 1.02 | - | - | - | - |
| 1977 | 0.21 | 1.01 | - | - | - | - |
| 1978 | 0.18 | 1.00 | - | - | - | - |
| 1979 | 0.20 | 1.00 | - | - | - | - |
| 1980 | 0.25 | 1.01 | - | - | - | - |
| 1981 | 0.23 | 1.00 | - | - | - | - |
| 1982 | 0.22 | 1.01 | - | - | - | - |
| 1983 | 0.30 | 1.01 | - | - | - | - |
| 1984 | 0.27 | 1.00 | - | - | - | - |


| 1985 | 0.37 | 1.02 | - | - | - | - |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1986 | 0.35 | 1.01 | - | - | - | - |
| 1987 | 0.39 | 1.01 | - | - | - | - |
| 1988 | 0.36 | 1.00 | - | - | - | - |
| 1989 | 0.28 | 1.01 | - | - | - | - |
| 1990 | 0.32 | 1.02 | - | - | - | - |
| 1991 | 0.27 | 1.01 | - | - | - | - |
| 1992 | 0.30 | 1.02 | - | - | - | - |
| 1993 | 0.29 | 1.02 | - | - | - | - |
| 1994 | 0.23 | 1.01 | - | - | - | - |
| 1995 | 0.20 | 1.01 | - | - | 8.33 | 2.12 |
| 1996 | 0.20 | 1.01 | - | - | 8.54 | 2.31 |
| 1997 | 0.14 | 1.02 | - | - | 9.18 | 2.05 |
| 1998 | 0.14 | 1.02 | - | - | 8.20 | 2.11 |
| 1999 | 0.17 | 1.01 | - | - | 11.20 | 1.46 |
| 2000 | 0.20 | 1.02 | 0.14 | 1.21 | 10.61 | 2.93 |
| 2001 | 0.24 | 1.04 | 0.17 | 1.15 | - | - |
| 2002 | 0.21 | 1.03 | 0.24 | 1.18 | - | - |
| 2003 | 0.16 | 1.01 | 0.19 | 1.11 | - | - |
| 2004 | 0.17 | 1.04 | 0.27 | 1.00 | - | - |
| 2005 | 0.18 | 1.04 | 0.17 | 1.00 | 13.33 | 1.14 |
| 2006 | 0.22 | 1.03 | 0.17 | 1.01 | 16.32 | 1.02 |
| 2007 | 0.18 | 1.05 | 0.16 | 1.03 | 13.83 | 1.18 |
| 2008 | 0.17 | 1.05 | 0.16 | 1.03 | 13.53 | 1.09 |
| 2009 | 0.20 | 1.07 | 0.16 | 1.06 | 10.90 | 1.23 |
| 2010 | 0.21 | 1.10 | 0.18 | 1.07 | 9.23 | 1.23 |
| 2011 | 0.17 | 1.08 | 0.16 | 1.04 | 11.70 | 1.00 |
| 2012 | 0.20 | 1.16 | 0.17 | 1.10 | 11.18 | 1.09 |

Table 2.2. Catch per unit effort (CPUE, \# of swordfish/1000 hooks) used for assessment of swordfish (Xiphias gladius) in the Eastern Pacific Ocean. A '-' indicates no effort or data available. Calculated relative CVs are reported here, but model runs assumed a relative CV of 1 for all values of all CPUE series.

| Year | Japan 1 <br> Deepwater \& offshore longline |  | Japan 2 <br> Deepwater \& offshore longline |  | Japan 3 <br> Deepwater \& offshore longline |  | Taiwan Deepwater longline |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Relative CV |  | Relative CV |  | Relative CV | CPUE | $\begin{aligned} & \text { Relative } \\ & \text { CV } \end{aligned}$ |
| 1951 | - | - | - | - | - | - | - | - |
| 1952 | - | - | - | - | - | - | - | - |
| 1953 | - | - | - | - | - | - | - | - |
| 1954 | - | - | - | - | - | - | - | - |
| 1955 | 0.07 | 1.29 | - | - | - | - | - | - |
| 1956 | 0.05 | 1.45 | - | - | - | - | - | - |
| 1957 | 0.20 | 1.08 | - | - | - | - | - | - |
| 1958 | 0.12 | 1.05 | - | - | - | - | - | - |
| 1959 | 0.07 | 1.05 | - | - | - | - | - | - |
| 1960 | 0.09 | 1.05 | - | - | - | - | - | - |
| 1961 | 0.16 | 1.02 | - | - | - | - | - | - |
| 1962 | 0.18 | 1.01 | - | - | - | - | - | - |
| 1963 | 0.23 | 1.00 | - | - | - | - | - | - |
| 1964 | 0.20 | 1.00 | - | - | - | - | - | - |
| 1965 | 0.17 | 1.00 | - | - | - | - | - | - |
| 1966 | 0.19 | 1.01 | - | - | - | - | - | - |
| 1967 | 0.20 | 1.01 | - | - | - | - | - | - |
| 1968 | 0.20 | 1.01 | - | - | - | - | - | - |
| 1969 | 0.24 | 1.01 | - | - | - | - | - | - |
| 1970 | 0.28 | 1.01 | - | - | - | - | - | - |
| 1971 | 0.22 | 1.01 | - | - | - | - | - | - |
| 1972 | 0.18 | 1.01 | - | - | - | - | - | - |
| 1973 | 0.25 | 1.01 | - | - | - | - | - | - |
| 1974 | 0.26 | 1.00 | - | - | - | - | - | - |
| 1975 | - | - | 0.35 | 2.01 | - | - | - | - |
| 1976 | - | - | 0.36 | 1.33 | - | - | - | - |
| 1977 | - | - | 0.39 | 2.65 | - | - | - | - |
| 1978 | - | - | 0.35 | 1.50 | - | - | - | - |
| 1979 | - | - | 0.29 | 1.12 | - | - | - | - |
| 1980 | - | - | 0.31 | 1.88 | - | - | - | - |
| 1981 | - | - | 0.38 | 3.10 | - | - | - | - |
| 1982 | - | - | 0.32 | 1.91 | - | - | - | - |
| 1983 | - | - | 0.32 | 1.76 | - | - | - | - |
| 1984 | - | - | 0.26 | 1.66 | - | - | - | - |


| 1985 | - | - | 0.24 | 1.15 | - | - | - | - |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1986 | - | - | 0.28 | 1.27 | - | - | - | - |
| 1987 | - | - | 0.30 | 1.41 | - | - | - | - |
| 1988 | - | - | 0.26 | 1.20 | - | - | - | - |
| 1989 | - | - | 0.26 | 1.14 | - | - | - | - |
| 1990 | - | - | 0.30 | 1.19 | - | - | - | - |
| 1991 | - | - | 0.26 | 1.00 | - | - | - | - |
| 1992 | - | - | 0.24 | 1.07 | - | - | - | - |
| 1993 | - | - | 0.27 | 1.14 | - | - | - | - |
| 1994 | - | - | - | - | 0.26 | 1.00 | - | - |
| 1995 | - | - | - | - | 0.27 | 1.01 | - | - |
| 1996 | - | - | - | - | 0.30 | 1.12 | - | - |
| 1997 | - | - | - | - | 0.35 | 1.39 | - | - |
| 1998 | - | - | - | - | 0.41 | 1.49 | - | - |
| 1999 | - | - | - | - | 0.39 | 1.46 | - | - |
| 2000 | - | - | - | - | 0.48 | 1.81 | 0.44 | 1.56 |
| 2001 | - | - | - | - | 0.55 | 2.08 | 0.57 | 1.00 |
| 2002 | - | - | - | - | 0.44 | 1.60 | 0.53 | 1.00 |
| 2003 | - | - | - | - | 0.41 | 1.50 | 0.50 | 1.01 |
| 2004 | - | - | - | - | 0.35 | 1.35 | 0.51 | 1.03 |
| 2005 | - | - | - | - | 0.35 | 1.41 | 0.43 | 1.04 |
| 2006 | - | - | - | - | 0.44 | 1.79 | 0.45 | 1.04 |
| 2007 | - | - | - | - | 0.52 | 2.01 | 0.48 | 1.14 |
| 2008 | - | - | - | - | 0.68 | 2.67 | 0.49 | 1.26 |
| 2009 | - | - | - | - | 0.85 | 3.34 | 0.57 | 1.29 |
| 2010 | - | - | - | - | 1.01 | 3.73 | 0.50 | 1.14 |
| 2011 | - | - | - | - | 1.00 | 3.86 | 0.51 | 1.23 |
| 2012 | - | - | - | - | 1.02 | 3.91 | 0.57 | 1.64 |

Table 3.1. Parameters and assumed prior distributions for a Bayesian generalized production model of swordfish (Xiphias gladius) in the Western and Central North Pacific Ocean.

| $\begin{gathered} \text { Paramete } \\ \mathbf{r} \end{gathered}$ | Description | Assumed Distribution | Assumed Mean | Assumed CV |
| :---: | :---: | :---: | :---: | :---: |
| $R$ | Intrinsic growth rate ( $\mathrm{yr}^{-1}$ ) | $R \sim \log N\left(\log (0.5)-\frac{\sigma_{R}^{2}}{2}, \sigma_{R}^{2}\right)$ | 0.5 | 50\% |
| K | Carrying capacity (1000 $\mathrm{mt})$ | $K \sim \log N\left(\log (150)-\frac{\sigma_{K}^{2}}{2}, \sigma_{K}^{2}\right)$ | $\begin{gathered} 150,000 \\ \mathrm{mt} \end{gathered}$ | 50\% |
| $S$ | Production shape parameter | $S \sim \operatorname{Gamma}(2,2)$ | 1.0 | 71\% |
| $Q$ | Catchability coefficient | $1 / Q \sim \operatorname{Gamma}(0.01,0.01)$ | $1 / Q=1.0$ | $\begin{gathered} \text { Variance }= \\ 1000 \end{gathered}$ |
| $P_{1}$ | Initial proportion of biomass to carrying capacity | $P_{1} \sim \log N\left(\log (0.9)-\frac{\sigma_{P_{1}}^{2}}{2}, \sigma_{P_{1}}^{2}\right)$ | 0.90 | 10\% |
| $\tau^{2}$ | Observation error variance | $1 / \tau^{2} \sim \operatorname{Gamma}(2,0.45)$ | 0.223 | 50\% |
| $\sigma^{2}$ | Process error variance | $1 / \sigma^{2} \sim \operatorname{Gamma}(4,0.1)$ | 0.025 | 16\% |

Table 3.2. Parameters and assumed prior distributions for a Bayesian generalized production model of swordfish (Xiphias gladius) in the Eastern Pacific Ocean.


Table 4.1. Summary of model diagnostics for a Bayesian state-space model of swordfish in the Western and Central North Pacific Ocean. DIC is the deviance information criterion (a model fit statistic used to compare models that use the same datasets), RMSE is root mean-squared error from fitted versus observed CPUE, and $\rho$ is the correlation coefficient between observed and predicted CPUE.

| Diagnostic | Index | Mean |
| :---: | :---: | :---: |
| DIC | Model | -185.49 |
| RMSE | Japan CPUE (1952-2012) | 0.033 |
| RMSE | Taiwan CPUE (2000-2012) | 0.038 |
| RMSE | Hawaii CPUE (1995-2012) | 2.273 |
| $\rho$ | Japan CPUE (1952-2012) | 0.80 |
| $\rho$ | Taiwan CPUE (2000-2012) | 0.18 |
| $\rho$ | Hawaii CPUE (1995-2012) | 0.41 |

Table 4.2. Summary of model diagnostics for a Bayesian state-space model of swordfish in the Eastern Pacific Ocean. DIC is the deviance information criterion (a model fit statistic used to compare models that use the same datasets), RMSE is root mean-squared error from fitted versus observed CPUE, and $\rho$ is the correlation coefficient between observed and predicted CPUE.

| Diagnostic | Index | Mean |
| :---: | :---: | :---: |
| DIC | Model | -112.54 |
| RMSE | Japan CPUE 1 (1952-1974) | 0.054 |
| RMSE | Japan CPUE 2 (1975-1993) | 0.052 |
| RMSE | Japan CPUE 3 (1994-2012) | 0.169 |
| RMSE | Taiwan CPUE (2000-2012) | 0.107 |
| $\rho$ | Japan CPUE 1 (1952-1974) | 0.61 |
| $\rho$ | Japan CPUE 2 (1975-1993) | 0.58 |
| $\rho$ | Japan CPUE 3 (1994-2012) | 0.79 |
| $\rho$ | Taiwan CPUE (2000-2012) | 0.41 |

Table 5.1. Estimated mean and standard deviation model parameter values for a Bayesian statespace production model for swordfish in the Western and Central North Pacific Ocean.
$\left.\begin{array}{ccc}\hline \text { Parameter } & \text { Mean } & \text { SD } \\ \hline \text { Intrinsic rate of pop. growth }(r) & 0.578 & 0.217 \\ \text { Carrying capacity }(K ; 1000 \mathrm{mt}) & 123.700 & 24.630 \\ \text { Production shape parameter }(M) & 0.978 & 0.453 \\ \text { Process error variance }\left(\sigma^{2}\right) & 0.017 & 0.005 \\ \text { JPN longline CPUE obs. error variance } \\ \left(\tau_{J P N}^{2}\right)\end{array} 0^{\left(P_{1}\left(B_{1951} / K\right)\right.} \begin{array}{c}\left(\tau_{T W N}^{2}\right) \\ \text { TWN longline error variance }\end{array}\right)$

Table 5.2. Estimated mean and standard deviation model parameter values for a Bayesian statespace production model for swordfish in the Eastern Pacific Ocean.

| Parameter | Mean | SD |
| :---: | :---: | :---: |
| Intrinsic rate of pop. growth (r) | 0.458 | 0.195 |
| Carrying capacity ( $K$; 1000 mt ) | 65.19 | 15.71 |
| Production shape parameter ( $M$ ) | 0.927 | 0.712 |
| Process error variance ( $\sigma^{2}$ ) | 0.030 | 0.011 |
| JPN longline CPUE I obs. error variance $\left(\tau_{J P N 1}^{2}\right)$ | 0.174 | 0.077 |
| JPN longline CPUE II obs. error variance $\left(\tau_{J P N 2}^{2}\right)$ | 0.069 | 0.026 |
| JPN longline CPUE III obs. error variance $\left(\tau_{J P N 3}^{2}\right)$ | 0.115 | 0.053 |
| TWN longline CPUE obs. error variance $\left(\tau_{T W N}^{2}\right)$ | 0.097 | 0.045 |
| $P_{1}\left(B_{1951} / K\right)$ | 0.880 | 0.089 |
| JPN longline CPUE I catchability ( $q_{J P N I}$ ) | $3.48 \times 10^{-3}$ | $1.31 \times 10^{-3}$ |
| JPN longline CPUE II catchability ( $q_{J P N 2}$ ) | $7.17 \times 10^{-3}$ | $2.23 \times 10^{-3}$ |
| JPN longline CPUE III catchability ( $q_{J P N 3}$ ) | 0.012 | $3.81 \times 10^{-3}$ |
| TWN longline CPUE III catchability ( $q_{J P N 3}$ ) | 0.011 | $3.69 \times 10^{-3}$ |
| Max surplus production (MSY; 1000 tons) | 5.49 | 1.63 |
| Biomass giving MSY ( $B_{\mathrm{MSY}} ; 1000$ tons $)$ | 31.17 | 6.99 |
| Harvest rate giving MSY ( $H_{\mathrm{MSY}}$ ) | 0.183 | 0.063 |

Table 6.1. Estimated mean values of exploitable biomass and harvest rate for swordfish in the Western and Central North Pacific Ocean.

| Year | Exploitable biomass $(1000 \mathrm{mt})$ |  | Harvest rate |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Mean | SD | Mean | SD |
| 1951 | 104.60 | 22.56 | 0.12 | 0.03 |
| 1952 | 87.01 | 20.48 | 0.14 | 0.03 |
| 1953 | 79.61 | 19.29 | 0.16 | 0.04 |
| 1954 | 81.83 | 19.85 | 0.18 | 0.04 |
| 1955 | 78.77 | 19.21 | 0.19 | 0.05 |
| 1956 | 74.86 | 18.18 | 0.22 | 0.05 |
| 1957 | 75.76 | 18.34 | 0.21 | 0.05 |
| 1958 | 81.88 | 19.45 | 0.25 | 0.06 |
| 1959 | 76.98 | 18.60 | 0.26 | 0.06 |
| 1960 | 77.21 | 18.51 | 0.30 | 0.07 |
| 1961 | 73.29 | 18.40 | 0.31 | 0.08 |
| 1962 | 68.93 | 18.20 | 0.19 | 0.05 |
| 1963 | 73.50 | 19.01 | 0.16 | 0.04 |
| 1964 | 74.68 | 19.10 | 0.10 | 0.03 |
| 1965 | 80.05 | 19.78 | 0.13 | 0.03 |
| 1966 | 78.82 | 19.31 | 0.15 | 0.04 |
| 1967 | 72.97 | 18.07 | 0.17 | 0.04 |
| 1968 | 68.54 | 17.12 | 0.17 | 0.04 |
| 1969 | 68.85 | 17.29 | 0.16 | 0.04 |
| 1970 | 70.63 | 17.82 | 0.14 | 0.03 |
| 1971 | 72.23 | 18.00 | 0.11 | 0.03 |
| 1972 | 74.62 | 18.45 | 0.10 | 0.02 |
| 1973 | 81.01 | 19.57 | 0.10 | 0.02 |
| 1974 | 86.16 | 20.67 | 0.10 | 0.02 |
| 1975 | 85.99 | 20.55 | 0.14 | 0.03 |
| 1976 | 85.85 | 20.82 | 0.15 | 0.04 |
| 1977 | 81.04 | 20.04 | 0.16 | 0.04 |
| 1978 | 77.99 | 19.31 | 0.19 | 0.05 |
| 1979 | 79.28 | 19.82 | 0.15 | 0.04 |
| 1980 | 85.73 | 21.33 | 0.13 | 0.03 |
| 1981 | 88.18 | 21.89 | 0.13 | 0.03 |
| 1982 | 91.10 | 22.48 | 0.11 | 0.03 |
| 1983 | 102.70 | 25.03 | 0.13 | 0.03 |
| 1984 | 106.30 | 26.40 | 0.13 | 0.03 |
| 1985 | 116.90 | 29.34 | 0.14 | 0.04 |
|  |  |  |  |  |


| 1986 | 117.70 | 30.05 | 0.12 | 0.03 |
| :--- | :--- | :--- | :--- | :--- |
| 1987 | 121.30 | 30.90 | 0.12 | 0.03 |
| 1988 | 116.70 | 29.96 | 0.12 | 0.03 |
| 1989 | 108.00 | 27.52 | 0.12 | 0.03 |
| 1990 | 108.70 | 27.16 | 0.12 | 0.03 |
| 1991 | 104.10 | 25.81 | 0.13 | 0.03 |
| 1992 | 103.80 | 25.30 | 0.19 | 0.05 |
| 1993 | 94.69 | 23.49 | 0.22 | 0.05 |
| 1994 | 80.19 | 20.43 | 0.20 | 0.05 |
| 1995 | 70.20 | 17.78 | 0.21 | 0.05 |
| 1996 | 65.65 | 16.37 | 0.21 | 0.05 |
| 1997 | 60.86 | 14.98 | 0.24 | 0.06 |
| 1998 | 60.20 | 14.81 | 0.24 | 0.06 |
| 1999 | 65.18 | 15.77 | 0.23 | 0.06 |
| 2000 | 69.82 | 16.65 | 0.26 | 0.06 |
| 2001 | 74.02 | 18.24 | 0.20 | 0.05 |
| 2002 | 75.47 | 18.48 | 0.19 | 0.05 |
| 2003 | 71.61 | 17.41 | 0.20 | 0.05 |
| 2004 | 73.37 | 17.81 | 0.19 | 0.05 |
| 2005 | 73.86 | 17.69 | 0.20 | 0.05 |
| 2006 | 76.32 | 18.32 | 0.21 | 0.05 |
| 2007 | 72.29 | 17.50 | 0.23 | 0.06 |
| 2008 | 68.62 | 16.85 | 0.21 | 0.05 |
| 2009 | 68.77 | 16.80 | 0.19 | 0.05 |
| 2010 | 68.97 | 16.95 | 0.16 | 0.04 |
| 2011 | 68.56 | 16.77 | 0.15 | 0.04 |
| 2012 | 72.50 | 17.50 | 0.14 | 0.03 |
|  |  |  |  |  |

Table 6.2. Estimated mean values of exploitable biomass and harvest rate for swordfish in the Eastern Pacific Ocean.

| Year | Exploitable biomass $(1000 \mathrm{mt})$ |  | Harvest rate |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Mean | SD | Mean | SD |
| 1951 | 57.21 | 14.46 | 0.00 | 0.00 |
| 1952 | 55.01 | 15.68 | 0.00 | 0.00 |
| 1953 | 51.47 | 16.98 | 0.00 | 0.00 |
| 1954 | 45.55 | 17.06 | 0.02 | 0.02 |
| 1955 | 38.36 | 14.45 | 0.01 | 0.01 |
| 1956 | 37.03 | 13.82 | 0.02 | 0.01 |
| 1957 | 41.69 | 14.47 | 0.02 | 0.01 |
| 1958 | 40.89 | 14.29 | 0.01 | 0.01 |
| 1959 | 39.68 | 14.15 | 0.01 | 0.01 |
| 1960 | 42.92 | 14.85 | 0.01 | 0.01 |
| 1961 | 49.29 | 16.16 | 0.02 | 0.01 |
| 1962 | 54.92 | 17.73 | 0.03 | 0.01 |
| 1963 | 59.32 | 19.06 | 0.03 | 0.01 |
| 1964 | 60.01 | 19.45 | 0.07 | 0.03 |
| 1965 | 58.08 | 19.45 | 0.04 | 0.02 |
| 1966 | 60.16 | 20.00 | 0.04 | 0.02 |
| 1967 | 62.37 | 20.70 | 0.05 | 0.02 |
| 1968 | 64.25 | 21.37 | 0.06 | 0.02 |
| 1969 | 67.07 | 22.31 | 0.12 | 0.04 |
| 1970 | 64.85 | 22.52 | 0.11 | 0.04 |
| 1971 | 60.99 | 21.47 | 0.04 | 0.02 |
| 1972 | 61.35 | 20.65 | 0.06 | 0.02 |
| 1973 | 62.89 | 20.72 | 0.10 | 0.04 |
| 1974 | 59.70 | 19.69 | 0.04 | 0.02 |
| 1975 | 56.52 | 17.91 | 0.04 | 0.01 |
| 1976 | 55.92 | 17.68 | 0.07 | 0.02 |
| 1977 | 54.95 | 17.63 | 0.06 | 0.02 |
| 1978 | 52.23 | 16.91 | 0.06 | 0.02 |
| 1979 | 49.50 | 16.15 | 0.04 | 0.01 |
| 1980 | 50.49 | 16.22 | 0.06 | 0.02 |
| 1981 | 51.96 | 16.69 | 0.08 | 0.03 |
| 1982 | 48.78 | 15.87 | 0.08 | 0.03 |
| 1983 | 46.16 | 15.19 | 0.05 | 0.02 |
| 1984 | 43.11 | 14.19 | 0.04 | 0.01 |
| 1985 | 42.18 | 13.84 | 0.03 | 0.01 |
|  |  |  |  |  |


| 1986 | 44.18 | 14.19 | 0.07 | 0.02 |
| :--- | :--- | :--- | :--- | :--- |
| 1987 | 44.52 | 14.26 | 0.09 | 0.03 |
| 1988 | 42.48 | 13.74 | 0.09 | 0.03 |
| 1989 | 42.19 | 13.63 | 0.08 | 0.03 |
| 1990 | 43.41 | 13.72 | 0.15 | 0.05 |
| 1991 | 39.84 | 12.97 | 0.12 | 0.04 |
| 1992 | 38.14 | 12.32 | 0.15 | 0.05 |
| 1993 | 35.86 | 11.82 | 0.12 | 0.04 |
| 1994 | 32.16 | 11.05 | 0.13 | 0.05 |
| 1995 | 31.51 | 11.07 | 0.11 | 0.04 |
| 1996 | 33.62 | 11.52 | 0.14 | 0.05 |
| 1997 | 36.17 | 12.18 | 0.18 | 0.06 |
| 1998 | 37.98 | 12.91 | 0.22 | 0.07 |
| 1999 | 37.90 | 13.49 | 0.10 | 0.04 |
| 2000 | 43.42 | 14.47 | 0.16 | 0.05 |
| 2001 | 46.46 | 15.61 | 0.18 | 0.06 |
| 2002 | 44.32 | 15.18 | 0.18 | 0.06 |
| 2003 | 41.98 | 14.62 | 0.15 | 0.05 |
| 2004 | 40.66 | 14.17 | 0.16 | 0.06 |
| 2005 | 39.55 | 13.94 | 0.10 | 0.03 |
| 2006 | 43.10 | 14.89 | 0.08 | 0.03 |
| 2007 | 47.98 | 16.29 | 0.09 | 0.03 |
| 2008 | 53.84 | 18.08 | 0.09 | 0.03 |
| 2009 | 60.57 | 20.29 | 0.14 | 0.04 |
| 2010 | 62.12 | 21.44 | 0.17 | 0.06 |
| 2011 | 60.81 | 21.83 | 0.18 | 0.06 |
| 2012 | 58.59 | 21.95 | 0.19 | 0.07 |
|  |  |  |  |  |
|  |  |  |  |  |

Table 7.1. Effects of high ( $+25 \%$ ) and low ( $-25 \%$ ) changes in prior means on model parameters including maximum sustainable yield, exploitable biomass to produce $M S Y$, and harvest rate to produce $M S Y$ for swordfish in the Western and Central North Pacific Ocean.

| Parameter | Base case |  | $1.25 * r$ |  | $0.75 * r$ |  | $1.25 * K$ |  | $0.75 * K$ |  | $1.25 * P_{1}$ |  | $0.75 * P_{1}$ |  | $1.25 * M$ |  | $0.75 * M$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | SD | \% change | CV | $\begin{gathered} \% \\ \text { change } \end{gathered}$ | CV | change | CV | $\begin{gathered} \% \\ \text { change } \end{gathered}$ | CV | $\begin{gathered} \% \\ \text { change } \end{gathered}$ | CV | change | CV | $\begin{gathered} \% \\ \text { change } \end{gathered}$ | CV | change | CV |
| $r$ | 0.58 | 0.22 | 13.79\% | 0.39 | -15.52\% | 0.37 | -1.72\% | 0.37 | 1.72\% | 0.37 | 0.00\% | 0.38 | -3.45\% | 0.38 | -5.17\% | 0.38 | 6.90\% | 0.37 |
| K | 123.66 | 24.63 | -1.91\% | 0.20 | 2.77\% | 0.20 | 3.52\% | 0.20 | -5.75\% | 0.20 | -8.18\% | 0.21 | 4.33\% | 0.20 | -1.16\% | 0.20 | 1.08\% | 0.20 |
| M | 0.98 | 0.45 | -11.22\% | 0.48 | 15.31\% | 0.44 | -1.02\% | 0.47 | 3.06\% | 0.46 | 17.35\% | 0.50 | -4.08\% | 0.48 | 9.18\% | 0.48 | -11.22\% | 0.44 |
| $P_{1}$ | 0.85 | 0.09 | 0.00\% | 0.11 | 0.00\% | 0.11 | 0.00\% | 0.11 | 0.00\% | 0.11 | 28.24\% | 0.12 | -23.53\% | 0.09 | 0.00\% | 0.11 | 0.00\% | 0.09 |
| $B_{\text {MSY }}$ | 60.72 | 11.79 | -4.05\% | 0.20 | 5.67\% | 0.20 | 3.28\% | 0.20 | -5.07\% | 0.19 | -5.57\% | 0.20 | 3.56\% | 0.20 | 0.41\% | 0.20 | -0.99\% | 0.19 |
| $B_{1951}$ | 104.60 | 22.37 | -1.72\% | 0.22 | 2.68\% | 0.22 | 3.35\% | 0.22 | -5.47\% | 0.21 | 17.69\% | 0.21 | -20.49\% | 0.22 | -0.96\% | 0.22 | 0.86\% | 0.21 |
| $B_{1951} / B_{\mathrm{MSY}}$ | 1.73 | 0.22 | 2.31\% | 0.12 | -2.89\% | 0.13 | 0.00\% | 0.13 | -0.58\% | 0.13 | 24.86\% | 0.13 | -23.12\% | 0.12 | -1.16\% | 0.13 | 1.73\% | 0.13 |
| $B_{2012}$ | 72.50 | 17.47 | -1.39\% | 0.25 | 2.33\% | 0.24 | 3.48\% | 0.24 | -5.41\% | 0.24 | -4.50\% | 0.25 | -2.22\% | 0.24 | -0.86\% | 0.24 | 0.58\% | 0.24 |
| $B_{2012} / B_{\mathrm{MSY}}$ | 1.20 | 0.19 | 2.50\% | 0.15 | -3.33\% | 0.16 | 0.00\% | 0.16 | -0.83\% | 0.16 | 0.83\% | 0.17 | -5.83\% | 0.16 | -1.67\% | 0.16 | 1.67\% | 0.16 |
| $H_{\text {MSY }}$ | 0.25 | 0.06 | 8.00\% | 0.22 | -4.00\% | 0.21 | 0.00\% | 0.24 | 8.00\% | 0.22 | 12.00\% | 0.21 | -4.00\% | 0.21 | 0.00\% | 0.24 | 4.00\% | 0.23 |
| $H_{1951}$ | 0.12 | 0.02 | 0.00\% | 0.25 | -8.33\% | 0.18 | -8.33\% | 0.18 | 0.00\% | 0.25 | -16.67\% | 0.20 | 25.00\% | 0.20 | 0.00\% | 0.25 | 0.00\% | 0.17 |
| $H_{1951} / H_{\text {MSY }}$ | 0.47 | 0.08 | -4.26\% | 0.18 | 4.26\% | 0.18 | 0.00\% | 0.17 | 0.00\% | 0.17 | -23.40\% | 0.19 | 31.91\% | 0.16 | 0.00\% | 0.19 | -2.13\% | 0.17 |
| $H_{2012}$ | 0.14 | 0.03 | 7.14\% | 0.27 | 0.00\% | 0.21 | 0.00\% | 0.21 | 7.14\% | 0.27 | 7.14\% | 0.27 | 7.14\% | 0.20 | 7.14\% | 0.20 | 0.00\% | 0.21 |
| $H_{2012} / H_{\text {MSY }}$ | 0.58 | 0.13 | -3.45\% | 0.23 | 5.17\% | 0.23 | 0.00\% | 0.22 | 0.00\% | 0.22 | -3.45\% | 0.23 | 8.62\% | 0.22 | 1.72\% | 0.22 | -1.72\% | 0.23 |
| MSY | 14.92 | 1.82 | 1.41\% | 0.12 | -1.68\% | 0.12 | 0.07\% | 0.12 | -0.20\% | 0.12 | 2.68\% | 0.12 | -2.61\% | 0.11 | -0.07\% | 0.12 | -0.07\% | 0.12 |

Table 7.2. Effects of high ( $+25 \%$ ) and low ( $-25 \%$ ) changes in prior means on model parameters including maximum sustainable yield, exploitable biomass to produce $M S Y$, and harvest rate to produce $M S Y$ for swordfish in the Eastern Pacific Ocean.

| Parameter | Base case |  | $1.25 * r$ |  | $0.75 * r$ |  | $1.25 * K$ |  | $0.75 * K$ |  | $1.25 * P_{1}$ |  | $0.75 * P_{1}$ |  | $1.25 * M$ |  | $0.75 * M$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | SD | \% change | CV | $\begin{gathered} \% \\ \text { change } \end{gathered}$ | CV | $\begin{gathered} \% \\ \text { change } \end{gathered}$ | CV | change | CV | $\begin{gathered} \% \\ \text { change } \end{gathered}$ | CV | $\begin{gathered} \% \\ \text { change } \end{gathered}$ | CV | change | CV | $\begin{gathered} \% \\ \text { change } \end{gathered}$ | CV |
| $r$ | 0.46 | 0.20 | 17.91\% | 0.44 | -19.31\% | 0.41 | -1.51\% | 0.43 | 2.29\% | 0.42 | 0.74\% | 0.43 | -3.86\% | 0.43 | -6.90\% | 0.44 | 7.05\% | 0.42 |
| K | 65.19 | 15.71 | -1.87\% | 0.24 | 2.91\% | 0.24 | 6.06\% | 0.24 | -7.49\% | 0.24 | -5.92\% | 0.25 | 7.79\% | 0.23 | -3.60\% | 0.25 | 2.84\% | 0.24 |
| M | 0.93 | 0.71 | -13.58\% | 0.77 | 19.62\% | 0.74 | -9.98\% | 0.68 | 7.02\% | 0.72 | 22.21\% | 0.78 | -15.87\% | 0.62 | 37.96\% | 0.85 | -21.83\% | 0.61 |
| $P_{1}$ | 0.88 | 0.09 | 0.08\% | 0.10 | -0.27\% | 0.10 | -0.33\% | 0.10 | 0.27\% | 0.10 | 26.07\% | 0.10 | -25.71\% | 0.10 | 0.40\% | 0.10 | -0.32\% | 0.10 |
| $B_{\text {MSY }}$ | 31.17 | 6.99 | -4.20\% | 0.22 | 6.26\% | 0.23 | 4.65\% | 0.23 | -6.45\% | 0.22 | -2.82\% | 0.24 | 4.17\% | 0.22 | 0.80\% | 0.24 | -0.48\% | 0.22 |
| $B_{1951}$ | 57.21 | 14.46 | -1.78\% | 0.25 | 2.66\% | 0.26 | 5.73\% | 0.26 | -7.22\% | 0.25 | 18.65\% | 0.26 | -20.89\% | 0.25 | -3.25\% | 0.27 | 2.53\% | 0.25 |
| $B_{1951} / B_{\mathrm{MSY}}$ | 1.84 | 0.26 | 2.44\% | 0.14 | -3.26\% | 0.14 | 0.98\% | 0.14 | -0.92\% | 0.14 | 22.27\% | 0.15 | -24.06\% | 0.13 | -3.86\% | 0.16 | 2.93\% | 0.13 |
| $B_{2012}$ | 58.59 | 21.95 | -2.92\% | 0.37 | 3.67\% | 0.38 | 6.57\% | 0.37 | -8.55\% | 0.38 | -7.63\% | 0.37 | 6.49\% | 0.37 | -4.92\% | 0.38 | 2.12\% | 0.37 |
| $B_{2012} / B_{\mathrm{MSY}}$ | 1.87 | 0.53 | 1.12\% | 0.28 | -1.87\% | 0.33 | 1.82\% | 0.28 | -2.51\% | 0.29 | -4.86\% | 0.28 | 2.46\% | 0.29 | -5.50\% | 0.29 | 2.72\% | 0.29 |
| $H_{\text {MSY }}$ | 0.18 | 0.06 | 8.20\% | 0.36 | -10.22\% | 0.35 | -5.19\% | 0.35 | 7.00\% | 0.35 | 9.95\% | 0.33 | -10.11\% | 0.35 | 4.16\% | 0.35 | -2.90\% | 0.35 |
| $H_{1951}$ | 0.00 | 0.00 | 0.00\% | 0.00 | 0.00\% | 0.00 | 0.00\% | 0.00 | 0.00\% | 0.00 | 0.00\% | 0.00 | 0.00\% | 0.00 | 0.00\% | 0.00 | 0.00\% | 0.00 |
| $H_{1951} / H_{\text {MSY }}$ | 0.00 | 0.00 | 0.00\% | 0.00 | 0.00\% | 0.00 | 0.00\% | 0.00 | 0.00\% | 0.00 | 0.00\% | 0.00 | 0.00\% | 0.00 | 0.00\% | 0.00 | 0.00\% | 0.00 |
| $\mathrm{H}_{2012}$ | 0.19 | 0.07 | 3.17\% | 0.38 | -3.37\% | 0.37 | -6.13\% | 0.37 | 9.61\% | 0.37 | 8.15\% | 0.37 | -6.28\% | 0.37 | 5.66\% | 0.38 | -2.18\% | 0.37 |
| $H_{2012} / H_{\text {MSY }}$ | 1.11 | 0.42 | -4.49\% | 0.37 | 8.62\% | 0.48 | -0.63\% | 0.39 | 2.06\% | 0.37 | -2.15\% | 0.37 | 4.67\% | 0.42 | 0.90\% | 0.36 | 1.35\% | 0.39 |
| MSY | 5.49 | 1.63 | 3.53\% | 0.31 | -4.57\% | 0.31 | -0.86\% | 0.30 | -0.09\% | 0.29 | 6.92\% | 0.30 | -6.60\% | 0.29 | 4.90\% | 0.31 | -3.59\% | 0.29 |

Table 8.1. Projected exploitable biomasses ( 1000 mt ) and harvest rates under eight different harvest scenarios during 2012-2016 for swordfish in the Western and Central North Pacific Ocean.

| Year | Recent harvest rate |  |  |  | Recent catch |  |  |  | Max obs harvest rate |  |  |  | FMSY |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Biomass |  | Harvest rate |  | Biomass |  | Harvest rate |  | Biomass |  | Harvest rate |  | Biomass |  | Harvest rate |  |
|  | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD |
| 2012 | 58.59 | 21.95 | 0.19 | 0.07 | 58.59 | 21.95 | 0.19 | 0.07 | 58.59 | 21.95 | 0.19 | 0.07 | 58.59 | 21.95 | 0.19 | 0.07 |
| 2013 | 50.40 | 21.69 | 0.18 | 0.07 | 50.37 | 21.56 | 0.23 | 0.10 | 50.35 | 21.76 | 0.24 | 0.08 | 50.40 | 21.73 | 0.19 | 0.07 |
| 2014 | 45.96 | 20.73 | 0.18 | 0.07 | 44.40 | 21.17 | 0.27 | 0.14 | 43.11 | 19.73 | 0.24 | 0.08 | 45.39 | 19.88 | 0.19 | 0.07 |
| 2015 | 42.89 | 19.85 | 0.18 | 0.07 | 39.09 | 20.98 | 0.33 | 0.19 | 38.51 | 18.32 | 0.24 | 0.08 | 42.27 | 18.69 | 0.19 | 0.07 |
| 2016 | 40.58 | 19.10 | 0.18 | 0.07 | 34.14 | 21.12 | 0.40 | 0.25 | 35.26 | 17.23 | 0.24 | 0.08 | 40.08 | 17.79 | 0.19 | 0.07 |


| Year | 0.5*FMSY |  |  |  | 0.75*FMSY |  |  |  | 1.25*FMSY |  |  |  | 1.5*FMSY |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Biomass |  | Harvest rate |  | Biomass |  | Harvest rate |  | Biomass |  | Harvest rate |  | Biomass |  | Harvest rate |  |
|  | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD |
| 2012 | 58.59 | 21.95 | 0.19 | 0.07 | 58.59 | 21.95 | 0.19 | 0.07 | 58.59 | 21.95 | 0.19 | 0.07 | 58.59 | 21.95 | 0.19 | 0.07 |
| 2013 | 50.32 | 21.62 | 0.10 | 0.04 | 50.35 | 21.66 | 0.14 | 0.06 | 50.32 | 21.58 | 0.23 | 0.08 | 50.34 | 21.68 | 0.27 | 0.09 |
| 2014 | 49.48 | 20.76 | 0.10 | 0.04 | 47.35 | 20.20 | 0.14 | 0.06 | 43.32 | 19.17 | 0.23 | 0.08 | 41.59 | 18.89 | 0.27 | 0.09 |
| 2015 | 48.96 | 20.30 | 0.10 | 0.04 | 45.40 | 19.42 | 0.14 | 0.06 | 39.22 | 17.72 | 0.23 | 0.08 | 36.67 | 17.19 | 0.27 | 0.09 |
| 2016 | 48.44 | 19.83 | 0.10 | 0.04 | 44.01 | 18.72 | 0.14 | 0.06 | 36.35 | 16.59 | 0.23 | 0.08 | 33.25 | 15.88 | 0.27 | 0.09 |


| Year | Recent harvest rate |  |  |  | Recent catch |  |  |  | Max obs harvest rate |  |  |  | FMSY |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Biomass |  | Harvest rate |  | Biomass |  | Harvest rate |  | Biomass |  | Harvest rate |  | Biomass |  | Harvest rate |  |
|  | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD |
| 2012 | 72.51 | 17.27 | 0.14 | 0.03 | 72.51 | 17.27 | 0.14 | 0.03 | 72.51 | 17.27 | 0.14 | 0.03 | 72.51 | 17.27 | 0.14 | 0.03 |
| 2013 | 77.09 | 20.06 | 0.15 | 0.04 | 77.17 | 20.05 | 0.14 | 0.04 | 77.27 | 20.17 | 0.32 | 0.08 | 77.16 | 20.1 | 0.25 | 0.06 |
| 2014 | 79.58 | 21.56 | 0.15 | 0.04 | 80.82 | 22.17 | 0.13 | 0.04 | 67.05 | 20.11 | 0.32 | 0.08 | 71.59 | 19.17 | 0.25 | 0.06 |
| 2015 | 81.19 | 22.24 | 0.15 | 0.04 | 83.68 | 23.66 | 0.13 | 0.04 | 61.18 | 19.86 | 0.32 | 0.08 | 68.4 | 18.75 | 0.25 | 0.06 |
| 2016 | 82.32 | 22.82 | 0.15 | 0.04 | 85.79 | 24.58 | 0.13 | 0.04 | 57.34 | 19.76 | 0.32 | 0.08 | 66.36 | 18.42 | 0.25 | 0.06 |


| Year | 0.5*FMSY |  |  |  | 0.75*FMSY |  |  |  | 1.25*FMSY |  |  |  | 1.5*FMSY |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Biomass |  | Harvest rate |  | Biomass |  | Harvest rate |  | Biomass |  | Harvest rate |  | Biomass |  | Harvest rate |  |
|  | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD |
| 2012 | 72.51 | 17.27 | 0.14 | 0.03 | 72.51 | 17.27 | 0.14 | 0.03 | 72.51 | 17.27 | 0.14 | 0.03 | 72.51 | 17.27 | 0.14 | 0.03 |
| 2013 | 77.14 | 20.05 | 0.14 | 0.03 | 77.13 | 20.09 | 0.2 | 0.05 | 77.14 | 20.15 | 0.3 | 0.07 | 77.11 | 19.99 | 0.35 | 0.07 |
| 2014 | 80.45 | 20.74 | 0.14 | 0.03 | 75.96 | 20 | 0.2 | 0.05 | 67.66 | 18.49 | 0.3 | 0.07 | 63.88 | 17.76 | 0.35 | 0.07 |
| 2015 | 83.04 | 21.36 | 0.14 | 0.03 | 75.32 | 19.97 | 0.2 | 0.05 | 62.38 | 17.61 | 0.3 | 0.07 | 56.94 | 16.52 | 0.35 | 0.07 |
| 2016 | 84.79 | 21.78 | 0.14 | 0.03 | 74.95 | 20.06 | 0.2 | 0.05 | 58.98 | 17.08 | 0.3 | 0.07 | 52.39 | 15.66 | 0.35 | 0.07 |

Table 8.2. Projected exploitable biomasses ( 1000 mt ) and harvest rates under eight different harvest scenarios during 2012-2016 for swordfish in the Eastern Pacific Ocean.

| Year | Recent harvest rate |  |  |  | Recent catch |  |  |  | Max obs harvest rate |  |  |  | FMSY |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Biomass |  | Harvest rate |  | Biomass |  | Harvest rate |  | Biomass |  | Harvest rate |  | Biomass |  | Harvest rate |  |
|  | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD |
| 2012 | 58.59 | 21.95 | 0.19 | 0.07 | 58.59 | 21.95 | 0.19 | 0.07 | 58.59 | 21.95 | 0.19 | 0.07 | 58.59 | 21.95 | 0.19 | 0.07 |
| 2013 | 50.40 | 21.69 | 0.18 | 0.07 | 50.37 | 21.56 | 0.23 | 0.10 | 50.35 | 21.76 | 0.24 | 0.08 | 50.40 | 21.73 | 0.19 | 0.07 |
| 2014 | 45.96 | 20.73 | 0.18 | 0.07 | 44.40 | 21.17 | 0.27 | 0.14 | 43.11 | 19.73 | 0.24 | 0.08 | 45.39 | 19.88 | 0.19 | 0.07 |
| 2015 | 42.89 | 19.85 | 0.18 | 0.07 | 39.09 | 20.98 | 0.33 | 0.19 | 38.51 | 18.32 | 0.24 | 0.08 | 42.27 | 18.69 | 0.19 | 0.07 |
| 2016 | 40.58 | 19.10 | 0.18 | 0.07 | 34.14 | 21.12 | 0.40 | 0.25 | 35.26 | 17.23 | 0.24 | 0.08 | 40.08 | 17.79 | 0.19 | 0.07 |
| Year | 0.5*FMSY |  |  |  | 0.75*FMSY |  |  |  | 1.25*FMSY |  |  |  | 1.5*FMSY |  |  |  |
|  | Biomass |  | Harvest rate |  | Biomass |  | Harvest rate |  | Biomass |  | Harvest rate |  | Biomass |  | Harvest rate |  |
|  | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD | Mean | SD |
| 2012 | 58.59 | 21.95 | 0.19 | 0.07 | 58.59 | 21.95 | 0.19 | 0.07 | 58.59 | 21.95 | 0.19 | 0.07 | 58.59 | 21.95 | 0.19 | 0.07 |
| 2013 | 50.32 | 21.62 | 0.10 | 0.04 | 50.35 | 21.66 | 0.14 | 0.06 | 50.32 | 21.58 | 0.23 | 0.08 | 50.34 | 21.68 | 0.27 | 0.09 |
| 2014 | 49.48 | 20.76 | 0.10 | 0.04 | 47.35 | 20.20 | 0.14 | 0.06 | 43.32 | 19.17 | 0.23 | 0.08 | 41.59 | 18.89 | 0.27 | 0.09 |
| 2015 | 48.96 | 20.30 | 0.10 | 0.04 | 45.40 | 19.42 | 0.14 | 0.06 | 39.22 | 17.72 | 0.23 | 0.08 | 36.67 | 17.19 | 0.27 | 0.09 |
| 2016 | 48.44 | 19.83 | 0.10 | 0.04 | 44.01 | 18.72 | 0.14 | 0.06 | 36.35 | 16.59 | 0.23 | 0.08 | 33.25 | 15.88 | 0.27 | 0.09 |

Table 9.1. Results from the final projected year of the risk analysis, 2016, for swordfish in the Western and Central North Pacific Ocean. Projected catch levels, probability of becoming overfished, and probability of overfishing are presented.

| Catch $(1000 \mathrm{mt})$ | Prob(B<0.5*BMSY) | Prob(H>HMSY) |
| :---: | :---: | :---: |
| 1.99 | 0.000 | 0.000 |
| 3.99 | 0.000 | 0.000 |
| 5.99 | 0.000 | 0.000 |
| 7.99 | 0.000 | 0.000 |
| 9.99 | 0.000 | 0.002 |
| 11.99 | 0.000 | 0.025 |
| 13.99 | 0.001 | 0.155 |
| 15.99 | 0.004 | 0.454 |
| 17.99 | 0.019 | 0.753 |
| 19.98 | 0.073 | 0.912 |
| 21.99 | 0.197 | 0.972 |
| 23.99 | 0.381 | 0.992 |
| 25.98 | 0.581 | 0.998 |
| 27.98 | 0.744 | 0.999 |
| 29.98 | 0.858 | 1.000 |
| 31.98 | 0.925 | 1.000 |
| 33.98 | 0.963 | 1.000 |
| 35.98 | 0.982 | 1.000 |
| 37.98 | 0.991 | 1.000 |
| 39.98 | 0.996 | 1.000 |
|  |  |  |

Table 9.2. Results from the final projected year of the risk analysis, 2016, for swordfish in the Eastern Pacific Ocean. Projected catch levels, probability of becoming overfished, and probability of overfishing are presented.

| Catch $(1000 \mathrm{mt})$ | Prob $\left(\mathrm{B}<0.5^{*} \mathrm{BMSY}\right)$ | Prob(H>HMSY) |
| :---: | :---: | :---: |
| 1.94 | 0.000 | 0.011 |
| 3.88 | 0.000 | 0.039 |
| 5.83 | 0.000 | 0.121 |
| 7.77 | 0.001 | 0.280 |
| 9.71 | 0.002 | 0.493 |
| 11.65 | 0.008 | 0.682 |
| 13.59 | 0.017 | 0.824 |
| 15.53 | 0.036 | 0.903 |
| 17.48 | 0.066 | 0.949 |
| 19.42 | 0.120 | 0.972 |
| 21.36 | 0.196 | 0.986 |
| 23.30 | 0.245 | 0.993 |
| 25.24 | 0.356 | 0.996 |
| 27.18 | 0.419 | 0.998 |
| 29.13 | 0.529 | 0.999 |
| 31.07 | 0.564 | 0.999 |
| 33.01 | 0.665 | 1.000 |
| 34.95 | 0.733 | 1.000 |
| 36.89 | 0.753 | 1.000 |
| 38.84 | 0.814 | 1.000 |



Figure 2. Two-stock structure for swordfish (Xiphias gladius) in the North Pacific Ocean, indicating separate stocks in the Western and Central Pacific Ocean and in the Eastern Pacific Ocean. This paper assesses swordfish in the Eastern Pacific Ocean.

fsFigure 2.1. Swordfish (Xiphias gladius) catch (metric tons) in the Western and Central North Pacific Ocean from 1951-2012 by country. $\dagger$ Other: catch data from Belize, Cook Islands, China, Spain, Fiji, Federated States of Micronesia, Kiribati, Marshall Islands, Papua New Guinea, Senegal, Tuvalu, and Vanuatu.


Figure 2.2. Swordfish (Xiphias gladius) catch (metric tons) in the Eastern Pacific Ocean from 1951-2012 by country. †Other: catch data from Belize, French Polynesia, United States, and Vanuatu.


Figure 3.1. Catch-per-unit-effort (CPUE) time series calculated from longline fisheries for swordfish (Xiphias gladius) in the Western and Central Pacific Ocean. The scale of CPUE for the Hawaii longline fishery is higher than the scales of the Japanese and Taiwanese CPUE series.


Figure 3.2. Catch-per-unit-effort (CPUE) time series calculated from longline fisheries for swordfish (Xiphias gladius) in the Eastern Pacific Ocean.


Figure 4.1. Bayesian state-space surplus production model predicted catch-per-unit-effort (CPUE) (dotted line, squares) and observed standardized CPUE (solid line, circles) for swordfish (Xiphias gladius) in the Western and Central North Pacific Ocean.


Figure 4.2. Bayesian state-space surplus production model predicted catch-per-unit-effort (CPUE) (dotted line, squares) and observed standardized CPUE (solid line, circles) for swordfish (Xiphias gladius) in the Eastern Pacific Ocean.


Figure 5.1. Trends in exploitable biomass (top) and harvest rate (bottom) of the Western and Central North Pacific Ocean swordfish, Xiphias gladius. Estimated mean values (black circles and solid line), $95 \%$ confidence interval bars, and estimated biological reference points ( $B_{M S Y}$ and $H_{M S Y}$, horizontal dashed lines) are presented.


Figure 5.2. Trends in exploitable biomass (top) and harvest rate (bottom) of the Eastern Pacific Ocean swordfish, Xiphias gladius. Estimated mean values (black circles and solid line), 95\% confidence interval bars, and estimated biological reference points ( $B_{M S Y}$ and $H_{M S Y}$, horizontal dashed lines) are presented.


Figure 6.1. Kobe diagram shows the estimated trajectories (1951-2012) of $B / B_{\mathrm{MSY}}$ and $H / H_{\mathrm{MSY}}$ for swordfish in the Western and Central North Pacific Ocean based on the base-case model.


Figure 6.2. Kobe diagram shows the estimated trajectories (1951-2012) of $B / B_{\mathrm{MSY}}$ and $H / H_{\mathrm{MSY}}$ for swordfish in the Eastern Pacific Ocean based on the base-case model.


Figure 7.1.1. Comparison of time-series of biomass estimates ( 1000 metric tons) from the basecase model with estimates from the models of different prior assumptions of intrinsic growth rate, $r$ (a), carrying capacity, $K(\mathrm{~b})$, initial condition, $P_{1}$ (c) and shape parameter, $M$ (d) for the swordfish in the Western and Central Pacific Ocean.


Figure 7.1.2. Comparison of time-series of harvest rates from the base-case model with estimates from the models of different prior assumptions of intrinsic growth rate, $r$ (a), carrying capacity, $K$ (b), initial condition, $P_{1}$ (c) and shape parameter, $M$ (d) for the swordfish in the Western and Central Pacific Ocean.


Figure 7.2.1. Comparison of time-series of biomass estimates ( 1000 metric tons) from the basecase model with estimates from the models of different prior assumptions of intrinsic growth rate, $r(\mathrm{a})$, carrying capacity, $K(\mathrm{~b})$, initial condition, $P_{1}$ (c) and shape parameter, $M$ (d) for the swordfish in the Eastern Pacific Ocean.


Figure 7.2.2. Comparison of time-series of harvest rates from the base-case model with estimates from the models of different prior assumptions of intrinsic growth rate, $r$ (a), carrying capacity, $K$ (b), initial condition, $P_{1}$ (c) and shape parameter, $M$ (d) for the swordfish in the Eastern Pacific Ocean.


Figure 8.1. Seven years within-model retrospective plots of the absolute change in biomass (a) and harvest rate (b) for the Western and Central North Pacific swordfish based on the base-case production model.


Figure 8.2. Seven years within-model retrospective plots of the absolute change in biomass (a) and harvest rate (b) for the Eastern Pacific swordfish based on the base-case production model.


Figure 9.1.1. Historic and 4 years projected trajectories of biomass ( 1000 metric tons) for swordfish in the Western and Central Pacific Ocean. Upper panel are fishing mortality at 0.5 , $0.75,1.0,1.25$, and $1.5 F_{\text {MSY }}$. Lower panel are status quo harvest rate from the most recent 3 years, status quo catch from the most recent 3 years and the maximum historically-observed harvest rate.


Figure 9.1.2. Historic and 4 years projected trajectories of harvest rate for swordfish in the Western and Central Pacific Ocean. Upper panel are fishing mortality at $0.5,0.75,1.0,1.25$, and $1.5 F_{\text {MSY }}$. Lower panel are status quo harvest rate from the most recent 3 years, status quo catch from the most recent 3 years and the maximum historically-observed harvest rate.


Figure 9.2.1. Historic and 4 years projected trajectories of biomass ( 1000 metric tons) for swordfish in the Eastern Pacific Ocean. Upper panel are fishing mortality at $0.5,0.75,1.0,1.25$, and $1.5 F_{\text {MSY }}$. Lower panel are status quo harvest rate from the most recent 3 years, status quo catch from the most recent 3 years and the maximum historically-observed harvest rate.


Figure 9.2.2. Historic and 4 years projected trajectories of harvest rate for swordfish in the Eastern Pacific Ocean. Upper panel are fishing mortality at $0.5,0.75,1.0,1.25$, and $1.5 F_{\text {MSY }}$. Lower panel are status quo harvest rate from the most recent 3 years, status quo catch from the most recent 3 years and the maximum historically-observed harvest rate.


Figure 10. Probabilities of experiencing overfishing ( $\mathrm{H}>H_{M S Y}$, solid line), of exploitable biomass falling below $B_{M S Y}\left(\mathrm{~B}<B_{M S Y}\right.$, open circles), and of being overfished relative to a reference level of $1 / 2 B_{M S Y}$ ( $\mathrm{B}<0.5^{*} B_{M S Y}$, solid squares) in 2016 for swordfish in the Western and Central Pacific Ocean stock area (a) and Eastern Pacific Ocean stock area (b) based on applying a constant catch biomass ( x -axis, thousand mt ) in the stock projections.

